# Contamination Assessment and Reduction Project Phase 2 (CARP II) 

Appendix A-2. CARP II Sampling and Analysis Program and Database

| Project Title: | NY/NJ Harbor Contamination Assessment and Reduction Project-CARP II |  |
| :--- | :--- | :---: |
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Chapter 1. Overview

This document summarizes the data collected under CARP II including field sampling and data analysis for: 1) Ambient Water, Head of Tide and Stormwater; 2) Harbor Wide Sediment Characterization; and 3) Navigation Channel and Off-Channel Characterization. Each of media were analyzed for congener PCBs (Method 1668) and the 17 toxic polychlorinated-dibenzo dioxins and furans (PCDD/Fs) by Method 1613. The PAHs, pesticides, and metals that were targets in CARP I were not sought in CARP II. The report also provides a data quality review of the PCDD/F and PCB data and information on how to access the data.

Water samples were collected with the TOPS device which produced filtered water ( 1 micron nominal pore size wound glass fiber cartridge filter), and suspended solids collected on the cartridge filter. The two phases were analyzed separately.

Sediment cores were divided into upper and lower portions which were analyzed separately. All of the grab sediment samples were analyzed for sediment characteristics. Worms (Neanthes virens) were exposed to sediments from all the core samples. All of these samples were tested for PCDD/Fs and congener PCBs. Also, accessory parameters were measured. All of the 42 grab samples were analyzed for sediment characteristics. Eight of them were also analyzed for PCDD/Fs and seven for congener PCBs.

A major focus of CARP II field work was to determine if there were differences in contaminant concentrations between navigational and non-navigational sediments. This is addressed in section 4.7. In summary, there were some differences but they were inconsistent and not very large.

## Chapter 2. Water

Water samples were collected using Trace Organics Platform Samplers (TOPS, Figure 1)). In total, 34 events were sampled with TOPS; seven Harbor ambient sites, five Head-of-Tide sites each visited twice; five storm water sites each visited twice, and seven Poughkeepsie sampling events (Figure 2). The TOPS samples were analyzed for dissolved and particle bound PCBs and PCDD/Fs, suspended solids, particulate organic carbon, and dissolved organic carbon.


Figure 1. Trace Organics Platform Sampler (TOPS). Here two units are collecting duplicate samples of suspended solids and filtered water. The carboy in the upper right is collecting filtered water to determine the volume of water passed through the glass fiber cartridge. Aqueous samples for chemical analysis are pumped directly into laboratory cleaned glass jugs. Power is being supplied by portable generator.

CARP II largely followed the procedures developed in CARP I for obtaining PCB and PCDD/F data from water samples by using TOPS (Trace Organics Platform Sampler). TOPS is a sampling platform designed to process large volumes of water on site to capture a sufficient mass of analyte for reliable detections of trace hydrophobic contaminants. TOPS was used in CARP I for both the dissolved phase (contaminants captured on the synthetic resin XAD) and the particulate phase. In CARP II the XAD resin was not used. Aqueous samples of filtered water were sent to the laboratory. Particles are captured on a 4 -inch long wound glass cartridge filter having a nominal porosity of 1 micron. Volumes of water filtered ranged from 17 to 1,800 liters depending on particle concentrations. TOPS automatically shuts off when back-pressure from the filter reaches 10 psi.

Operation of the TOPS requires two considerations. Estuary ambient waters may contain appreciable concentrations of large zooplankton that may complicate interpretation of the data. Therefore a Nytex 110 micron porosity plankton net was used to exclude large zooplankton. The second consideration is that the 1 micron glass cartridge filter permits some blow-by of small particles. This is corrected by calculating the trapping efficiency through measuring pre- and post-filtration with particulate organic
carbon (POC) captured on 0.7 micron flat glass filters from influent (inf) and effluent (eff) water from the glass fiber cartridge filter. Raw concentrations were divided by a POC efficiency calculated as (1-eff/inf).

PCBs were calculated as the sum of the aqueous and particulate phases. PCDD/Fs were assumed to be non-detectable in the aqueous phase and only analyzed from the particulate phase.

Site name abbreviations used in the graphs and tables of this report are shown in Table 1.

Table 1 Site name abbreviations used in the graphs and tables of this report

| site_type | Site name | Abbr. |
| :--- | :--- | :--- |
| Ambient | Arthur Kill | AK |
| Ambient | Hackensack River, ambient | Hra |
| Ambient | Kill van Kull | KVK |
| Ambient | Lower Bay | LB |
| Ambient | Newark Bay | NB |
| Ambient | Raritan Bay | RB |
| Ambient | Upper Bay | UB |
| Head_of_Tide | Dundee Dam, Passaic River | Dun |
| Head_of_Tide | Elizabeth River | ER |
| Head_of_Tide | Hackensack River | HR |
| Head_of_Tide | Passaic River | PR |
| Head_of_Tide | Raritan River | RR |
| Head_of_Tide | Saddle River | SR |
| Poughkeepsie | HRECOS pump station | Pough |
| Poughkeepsie | Poughkeepsie Drinking Water | Pough |
| Storm_Water | Bayonne | Bay |
| Storm_Water | Belleville | Bell |
| Storm_Water | Kearny | Kea |
| Storm_Water | Keyport Waterfront | Key |
| Storm_Water | New Milford | NM |



Figure 2. TOPS sampling sites for water.

### 2.1. Accessory Parameters

Accessory parameters particulate organic carbon (POC) and total suspended solids (TSS) were measured at the TOPS sites (Table 1).

Table 2 Accessory Parameters measured at TOPS sites.

|  | POC | TSS |  | POC | TSS |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Ambient |  |  | Poughkeepsie |  |  |
| AK-07/26/19 | 0.389 | 5.5 | Pough-01/09/19 | 1.4 | 17.9 |
| Hra-07/16/19 | 1.14 | 19 | Pough-01/14/19 | 1.12 | 16.4 |
| KVK-07/26/19 | 0.31 | 8.3 | Pough-03/13/19 | 0.373 | 41.3 |
| LB-08/07/19 | 0.588 | 16.7 | Pough-03/18/19 | 0.971 | 70 |
| NB-07/16/19 | 0.837 | 11.3 | Pough-04/22/19 |  | 12.9 |
| RB-08/22/19 | 1.22 | 8 | Pough-06/05/19 | 2.17 | 43.8 |
| UB-08/07/19 | 0.513 | 16.7 | Pough-12/11/18 <br> Storm_Water | 1.84 | 31.6 |
| Head_of_Tide |  |  |  |  |  |


| Dun-01/15/19 | 0.641 | ND | Bay-06/10/19 | 1.44 | ND |
| :--- | :---: | :---: | :--- | :---: | :---: |
| Dun-09/20/19 | 0.357 | ND | Bay-10/16/19 | 2.6 | 4.4 |
| ER-02/08/19 | 3.28 | 14.8 | Bell-07/23/19 | 40.1 | 114 |
| ER-12/25/19 | 1.05 | ND | Bell-10/03/19 | 41.3 | 154 |
| HR-02/26/19 | 1.25 | 4.2 | Kea-07/23/19 | 2.94 | 12.1 |
| HR-12/13/19 | 0.723 | ND | Kea-12/02/19 | 7.75 | 20 |
| RR-01/09/19 | 0.895 | 9.3 | Key-05/28/19 | 3.21 | 15.6 |
| RR-12/27/19 | 0.621 | ND | Key-12/09/19 | 6.14 | 9.3 |
| SR-01/15/19 | 0.669 | ND | NM-06/10/19 | 10.5 | 57.7 |
| SR-09/26/19 | 0.425 | ND | NM-12/02/19 | 7.17 | 23.3 |

### 2.2. Ambient Water - PCDD/Fs and PCBs

For ease of presentation the PCDD/F congeners were assigned an Order. The order value (from 1-17) will be used in the subsequent tables and graphs instead of an unwieldy chemical name. Toxic Equivalency Factors (TEFs) relate the toxicity of each PCDD/F congener to that of 2,3,7,8-TCDD. By multiplying the concentrations of each congener by its TEF a Toxic Equivalency (TEQ) is obtains. TEQs can meaningfully be summed.

Table 3. World Health Organization Toxic Equivalency Factors (TEFs) for PCDD/Fs.

| Analyte Name | TEF WHO 2005 | Order |
| :--- | :---: | :---: |
| $2,3,7,8-$ TCDD | 1 | 1 |
| $1,2,3,7,8-$ PECDD | 1 | 2 |
| $1,2,3,4,7,8-$ HXCDD | 0.1 | 3 |
| $1,2,3,6,7,8-$ HXCDD | 0.1 | 4 |
| $1,2,3,7,8,9-$ HXCDD | 0.1 | 5 |
| $1,2,3,4,6,7,8-$ HPCDD | 0.01 | 6 |
| OCDD | 0.0003 | 7 |
| 2,3,7,8-TCDF | 0.1 | 8 |
| $1,2,3,7,8-$ PECDF | 0.03 | 9 |
| $2,3,4,7,8-$ PECDF | 0.3 | 10 |
| $1,2,3,6,7,8-H X C D F$ | 0.1 | 11 |
| $1,2,3,4,7,8-$ HXCDF | 0.1 | 12 |
| $1,2,3,7,8,9-$ HXCDF | 0.1 | 13 |
| $2,3,4,6,7,8-H X C D F$ | 0.1 | 14 |
| $1,2,3,4,6,7,8-$ HPCDF | 0.01 | 15 |
| $1,2,3,4,7,8,9-H P C D F$ | 0.01 | 16 |
| OCDF | 0.0003 | 17 |

Similarly, PCBs are displayed as homologs. The ten homologs (1-10) encompass congeners of the same molecular weight. Homolog \#1 has one chlorine whereas homolog \#10 has ten chlorines. State
and Federal regulations assume that all PCBs have the same toxicity so they can be summed without needing toxicity factors. This subject is discussed in Section 5.

Total TEQ in all the CARP II ambient samples were dominated by 2,3,7,8-TCDD (congener \#1) but the magnitude of the dominance decreased with increasing distance from the Passaic River. 2,3,7,8-TCDD dominance was greatest, and virtually identical, in samples from Newark Bay (NB), Kill van Kull (KVK), and the Hackensack River (HR). The relative contribution of 1,2,3,4,6,7,8-HPCDD (congener \# 6), was relatively minor despite its prominence in the Hudson River samples taken at Poughkeepsie.
$1,2,3,7,8-$ PECDD (\#2) was, unusually, more abundant than 2,3,7,8-TCDD at Raritan Bay (RB). 1,2,3,7,8PECDD was also abundant in the Keyport storm water samples. Table 3 and Figure 3 show the PCDD/F congener concentrations (as $\mathrm{fg} / \mathrm{L}$ TEQ).

Table 4. PCDD/Fs in TOPS samples, TEQ in fg/L. See Table 3 for chemical names of the orders.

| Order | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | sum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ambient |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| AK-07/26/19 | 188 | 43 | 6 | 19 | 24 | 46 | 20 | 50 | 3 | 44 | 19 | 40 | 2 | 11 | 28 | 2 | 1 | 549 |
| Hra-07/16/19 | 1254 | 84 | 8 | 32 | 26 | 60 | 25 | 111 | 5 | 92 | 60 | 240 | 1 | 25 | 115 | 3 | 5 | 2,146 |
| KVK-07/26/19 | 209 | 0 | 2 | 12 | 14 | 23 | 8 | 31 | 2 | 22 | 9 | 22 | 1 | 6 | 16 | 1 | 1 | 376 |
| LB-08/07/19 | 83 | 32 | 5 | 18 | 16 | 28 | 9 | 33 | 2 | 26 | 7 | 13 | 0 | 5 | 9 | 0 | 0 | 288 |
| NB-07/16/19 | 448 | 34 | 5 | 15 | 11 | 28 | 9 | 53 | 3 | 26 | 17 | 59 | 0 | 9 | 32 | 1 | 1 | 753 |
| RB-08/22/19 | 8 | 14 | 1 | 4 | 4 | 7 | 3 | 9 | 0 | 5 | 2 | 3 | 1 | 2 | 3 | 0 | 0 | 67 |
| UB-08/07/19 | 50 | 32 | 4 | 15 | 17 | 30 | 9 | 38 | 2 | 29 | 7 | 13 | 0 | 5 | 10 | 1 | 0 | 263 |
| Head_of_Tide |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Dun-01/15/19 | 15 | 18 | 3 | 7 | 8 | 21 | 8 | 4 | 0 | 4 | 4 | 5 | 0 | 3 | 9 | 0 | 0 | 110 |
| Dun-09/20/19 | 10 | 18 | 8 | 22 | 21 | 75 | 25 | 16 | 1 | 22 | 15 | 16 | 0 | 11 | 29 | 1 | 2 | 291 |
| ER-02/08/19 | 23 | 88 | 15 | 41 | 32 | 116 | 44 | 17 | 2 | 26 | 24 | 29 | 1 | 18 | 51 | 3 | 4 | 532 |
| ER-12/25/19 | 0 | 0 | 2 | 4 | 5 | 13 | 6 | 3 | 0 | 5 | 4 | 6 | 0 | 3 | 6 | 0 | 0 | 58 |
| HR-12/13/19 | 5 | 18 | 2 | 5 | 7 | 14 | 5 | 4 | 0 | 4 | 4 | 4 | 0 | 3 | 6 | 0 | 0 | 82 |
| RR-01/09/19 | 2 | 7 | 2 | 5 | 5 | 20 | 27 | 2 | 0 | 3 | 2 | 3 | 0 | 1 | 4 | 0 | 0 | 85 |
| RR-12/27/19 | 0 | 9 | 3 | 7 | 8 | 29 | 39 | 3 | 0 | 5 | 3 | 3 | 0 | 3 | 4 | 0 | 0 | 116 |
| SR-01/15/19 | 7 | 10 | 3 | 7 | 7 | 19 | 6 | 3 | 0 | 4 | 4 | 4 | 0 | 3 | 8 | 0 | 1 | 85 |
| SR-09/26/19 | 46 | 247 | 29 | 76 | 87 | 117 | 19 | 14 | 1 | 23 | 22 | 19 | 0 | 11 | 31 | 3 | 1 | 747 |
| Poughkeepsie |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1/9/2019 | 6 | 19 | 3 | 13 | 11 | 33 | 12 | 16 | 1 | 13 | 4 | 5 | 0 | 2 | 7 | 0 | 0 | 145 |
| 1/14/2019 | 3 | 4 | 1 | 3 | 3 | 9 | 3 | 4 | 0 | 3 | 1 | 1 | 0 | 1 | 2 | 0 | 0 | 38 |
| 3/13/2019 | 12 | 23 | 8 | 26 | 13 | 80 | 24 | 28 | 1 | 19 | 7 | 7 | 1 | 4 | 11 | 1 | 1 | 265 |
| 3/18/2019 | 9 | 28 | 6 | 21 | 16 | 55 | 17 | 30 | 1 | 15 | 5 | 6 | 0 | 3 | 9 | 0 | 0 | 222 |
| 4/22/2019 | 3 | 5 | 1 | 3 | 2 | 7 | 3 | 3 | 0 | 3 | 1 | 1 | 0 | 1 | 2 | 0 | 0 | 36 |
| 6/5/2019 | 23 | 60 | 14 | 44 | 35 | 133 | 42 | 54 | 2 | 39 | 12 | 14 | 1 | 7 | 20 | 1 | 1 | 502 |
| 12/11/2018 | 7 | 20 | 3 | 11 | 9 | 26 | 9 | 13 | 1 | 13 | 3 | 4 | 1 | 3 | 6 | 0 | 0 | 131 |
| Storm_Water |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Bay-06/10/19 | 0 | 546 | 88 | 253 | 250 | 476 | 116 | 144 | 16 | 154 | 68 | 108 | 19 | 55 | 81 | 6 | 2 | 2,382 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bay-10/16/19 | 85 | 223 | 33 | 74 | 88 | 229 | 55 | 47 | 3 | 54 | 25 | 41 | 1 | 21 | 34 | 2 | 2 | 1,017 |
| Bell-07/23/19 | 75 | 102 | 25 | 150 | 77 | 537 | 129 | 27 | 3 | 40 | 27 | 45 | 2 | 19 | 40 | 2 | 2 | 1,305 |
| Bell-10/03/19 | 281 | 569 | 52 | 252 | 182 | 690 | 256 | 149 | 10 | 168 | 88 | 139 | 1 | 67 | 98 | 6 | 5 | 3,013 |
| Kea-07/23/19 | 41 | 141 | 40 | 151 | 116 | 501 | 87 | 18 | 2 | 37 | 47 | 66 | 1 | 34 | 96 | 4 | 4 | 1,385 |
| Key-05/28/19 | 153 | 1082 | 187 | 510 | 456 | 1473 | 298 | 50 | 6 | 81 | 140 | 102 | 2 | 107 | 236 | 12 | 11 | 4,907 |
| Key-12/09/19 | 332 | 4163 | 921 | 2901 | 2245 | 7720 | 1870 | 76 | 23 | 225 | 787 | 477 | 13 | 545 | 1426 | 70 | 69 | 23,864 |
| NM-06/10/19 | 391 | 759 | 207 | 830 | 515 | 4011 | 747 | 35 | 8 | 97 | 381 | 294 | 6 | 254 | 977 | 55 | 69 | 9,636 |
| NM-12/02/19 | 97 | 1087 | 327 | 1468 | 836 | 7899 | 1336 | 28 | 10 | 101 | 474 | 415 | 9 | 249 | 1199 | 73 | 109 | 15,717 |



Figure 5. Relative Abundance of PCDD/F TEQ in ambient suspended sediment.

Water data for PCBs are shown below summed by homologs (1-10) and as totals.

Table 5 PCBs in TOPS samples, pg/L.

| Homologs | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | sum |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ambient |  |  |  |  |  |  |  |  |  |  |  |
| AK-07/26/19 | 0.019 | 0.57 | 2.033 | 2.774 | 1.677 | 1.067 | 0.549 | 0.181 | 0.051 | 0.035 | 8.95 |
| Hra-07/16/19 | 0.039 | 0.745 | 2.74 | 4.719 | 2.507 | 1.309 | 0.621 | 0.227 | 0.075 | 0.053 | 13.03 |
| KVK-07/26/19 | 0.032 | 0.597 | 1.397 | 1.544 | 0.803 | 0.479 | 0.211 | 0.086 | 0.04 | 0.019 | 5.21 |
| LB-08/07/19 | 0.074 | 0.717 | 1.567 | 1.58 | 0.756 | 0.465 | 0.21 | 0.084 | 0.037 | 0.024 | 5.51 |
| NB-07/16/19 | 0.031 | 0.59 | 1.561 | 1.961 | 1.093 | 0.616 | 0.277 | 0.101 | 0.036 | 0.023 | 6.29 |
| RB-08/22/19 | 0.018 | 0.22 | 0.518 | 0.63 | 0.354 | 0.191 | 0.078 | 0.024 | 0.011 | 0.008 | 2.05 |
| UB-08/07/19 | 0.066 | 0.72 | 1.669 | 1.777 | 0.98 | 0.652 | 0.307 | 0.116 | 0.05 | 0.034 | 6.37 |
| Head of Tide |  |  |  |  |  |  |  |  |  |  |  |


| Dun-01/15/19 | 0 | 0.069 | 0.071 | 0.104 | 0.098 | 0.076 | 0.031 | 0.011 | 0.004 | 0.002 | 0.47 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Dun-09/20/19 | 0.004 | 0.1 | 0.312 | 0.598 | 0.628 | 0.567 | 0.254 | 0.117 | 0.052 | 0.031 | 2.66 |
| ER-02/08/19 | 0.006 | 0.197 | 0.325 | 0.579 | 1.102 | 1.734 | 1.196 | 0.341 | 0.075 | 0.029 | 5.58 |
| ER-12/25/19 | 0.027 | 0.365 | 0.458 | 0.375 | 0.358 | 0.612 | 0.389 | 0.107 | 0.013 | 0.005 | 2.71 |
| HR-02/26/19 | 0.006 | 0.051 | 0.035 | 0.037 | 0.057 | 0.054 | 0.022 | 0.006 | 0.001 | 0.001 | 0.27 |
| HR-12/13/19 | 0.002 | 0.067 | 0.047 | 0.084 | 0.138 | 0.11 | 0.046 | 0.015 | 0.003 | 0.002 | 0.51 |
| RR-01/09/19 | 0 | 0.017 | 0.033 | 0.055 | 0.091 | 0.109 | 0.063 | 0.019 | 0.006 | 0.002 | 0.4 |
| RR-12/27/19 | 0.005 | 0.038 | 0.04 | 0.067 | 0.116 | 0.13 | 0.066 | 0.024 | 0.006 | 0.003 | 0.49 |
| SR-01/15/19 | 0.005 | 0.073 | 0.1 | 0.089 | 0.083 | 0.063 | 0.024 | 0.006 | 0.002 | 0.001 | 0.45 |
| SR-09/26/19 | 0.217 | 0.17 | 0.362 | 0.441 | 0.514 | 0.514 | 0.159 | 0.062 | 0.013 | 0.005 | 2.46 |
| Poughkeepsie |  |  |  |  |  |  |  |  |  |  |  |
| Pough-01/09/19 | 0.185 | 1.602 | 3.288 | 2.757 | 0.74 | 0.328 | 0.09 | 0.026 | 0.013 | 0.007 | 9.03 |
| Pough-01/14/19 | 0.116 | 1.034 | 1.886 | 1.421 | 0.511 | 0.227 | 0.064 | 0.019 | 0.009 | 0.004 | 5.29 |
| Pough-03/13/19 | 0.774 | 4.962 | 9.15 | 6.217 | 2.259 | 1.119 | 0.335 | 0.133 | 0.065 | 0.029 | 25.04 |
| Pough-03/18/19 | 0.832 | 5.596 | 10.252 | 6.359 | 2.239 | 1.041 | 0.311 | 0.161 | 0.132 | 0.039 | 26.96 |
| Pough-04/22/19 | 0.196 | 2.245 | 3.27 | 2.26 | 0.732 | 0.352 | 0.093 | 0.031 | 0.015 | 0.007 | 9.2 |
| Pough-06/05/19 | 1.472 | 9.036 | 17.51 | 12.977 | 4.798 | 2.136 | 0.641 | 0.233 | 0.114 | 0.055 | 48.97 |
| Pough-12/11/18 | 0.21 | 1.993 | 3.458 | 2.686 | 0.928 | 0.388 | 0.099 | 0.029 | 0.014 | 0.007 | 9.81 |
| Storm Water |  |  |  |  |  |  |  |  |  |  |  |
| Bay-06/10/19 | 0.068 | 1.436 | 1.763 | 2.485 | 4.431 | 4.325 | 1.885 | 0.715 | 0.179 | 0.065 | 17.35 |
| Bay-10/16/19 | 0.014 | 0.445 | 0.393 | 1.096 | 3.177 | 2.477 | 1.025 | 0.381 | 0.092 | 0.035 | 9.14 |
| Bell-07/23/19 | 0.064 | 0.622 | 1.15 | 3.415 | 8.973 | 6.792 | 2.351 | 0.769 | 0.249 | 0.259 | 24.64 |
| Bell-10/03/19 | 0.135 | 1.548 | 1.976 | 5.665 | 15.71 | 22.144 | 21.87 | 8.342 | 0.94 | 0.401 | 78.73 |
| Kea-07/23/19 | 0.001 | 0.083 | 0.097 | 0.409 | 1.001 | 1.121 | 0.523 | 0.193 | 0.073 | 0.065 | 3.57 |
| Kea-12/02/19 | 0.016 | 0.909 | 0.451 | 0.924 | 1.578 | 1.787 | 0.896 | 0.328 | 0.097 | 0.053 | 7.04 |
| Key-05/28/19 | 0.013 | 1.162 | 0.216 | 0.388 | 1.009 | 1.242 | 0.49 | 0.157 | 0.044 | 0.023 | 4.74 |
| Key-12/09/19 | 0.012 | 1.546 | 0.461 | 0.9 | 1.898 | 1.758 | 0.779 | 0.274 | 0.082 | 0.034 | 7.75 |
| NM-06/10/19 | 0.032 | 3.045 | 1.912 | 1.394 | 1.047 | 0.921 | 0.308 | 0.095 | 0.035 | 0.018 | 8.81 |
| NM-12/02/19 | 0.102 | 4.274 | 1.095 | 0.62 | 0.801 | 0.796 | 0.335 | 0.106 | 0.028 | 0.012 | 8.17 |



Figure 6. Total PCB homolog concentrations in water, suspended sediment and filtered water.

### 2.4. Head of Tide PCDD/Fs and PCBs

2,3,7,8-TCDD was never dominant in any of the HOT or storm water samples. Total TEQ at the HOT Passaic River site, the Dundee Dam, were both low. Discharges were 1845 CFS on 1.15/19 and 231 on 9/20/19.


Figure 7. Relative Abundance of PCDD/F TEQ at Dundee Dam in suspended sediment.


Figure 8. Relative Abundance of PCB homologs at Dundee Dam, suspended sediment and filtered water.

The Elizabeth River showed one of the higher TEQs (average $0.532 \mathrm{pg} / \mathrm{L}$ ) on 2/8/19 (57 CFS) and one of the lowest TEQs ( $0.06 \mathrm{pg} / \mathrm{L}$ ) on 12/25/19 (15 CFS).


Figure 9. Relative Abundance of PCDD/F TEQ at Elizabeth River Site, suspended sediment.


Figure 10. Relative Abundance of PCB homologs at Elizbeth River Site, suspended sediment and filtered water.

The HOT Hackensack was sampled on 2/26/19 and on 12/13/19 but the PCDD/F sample from 2/26/19 was lost in a lab accident. Discharge on $2 / 26 / 19$ was 213 CFS but only 2.15 CFS on $12 / 13 / 19$.


Figure 11. Relative Abundance of PCDD/F TEQ at Hackensack River Site, suspended sediment.


Figure 12. Relative Abundance of PCB homologs at Hackensack River Site, suspended sediment and filtered water.

HOT Raritan River (at Bound Brook, NJ) was sampled on 1/9/19 (2487 CFS) and on 12/27/19 (722 CFS). This was only site where OCDD (\# 7) was the dominant contributor to total TEQ. The NOAA historical sediment database compiled under CARP II project has a similar pattern of OCDD dominance at the Tremley Point on the Arthur Kill.


Figure 13. Relative Abundance of PCDD/F TEQ at Raritan River Site, suspended sediment.


Figure 14. Relative Abundance of PCB homologs at Raritan River, suspended sediment and filtered water.

The Saddle River HOT samples ( $1 / 15 / 19$; 123 CFS and $9 / 26 / 19 ; 25.9$ CFS) is another example, like the Passaic at the Dundee, of apparent dilution where concentrations were higher at lower discharges. The patterns of the congener abundances were, like the Elizabeth River, distinctive.


Figure 25. Relative Abundance of PCDD/F TEQ at Saddle River Site, suspended sediment.


Figure 16. Relative Abundance of PCB homologs at Saddle River Site, suspended sediment and filtered water.

### 2.4. Poughkeepsie PCDD/Fs and PCBs

2,3,7,8-TCDD, so prominent in the harbor, was a minor component of the total TEQ in the Poughkeepsie samples (Figure 15). The patterns of relative abundances at all Poughkeepsie events were very similar. The hydrological regime of the Hudson River at Poughkeepsie is dominated by the tidal cycles. The graphs in Figure 16 indicate the times of sampling (in red) in the tidal cycle at Poughkeepsie.


Figure 17. Relative abundances of PCDD/F TEQ at Poughkeepsie, suspended sediment.


Figure 38. Relative Abundance of PCB homologs at Poughkeepsie, suspended sediment and filtered water.

Hudson Estuary suspended sediment concentrations are strongly affected by tides. A plot of river elevation and NTU, a surrogate for suspended sediment concentration, shows concentrations dramatically rising on the rising limb of the tidal cycle.


Figure 19. Hudson River turbidity (NTU) and surface elevation at Norrie Point. HRECOS data.
TOPS sampling occurs over several hours. None of the Poughkeepsie sampling events occurred during the rising limb of the hydrograph so we could expect to have seen much higher contaminant concentrations had the samples been taken at different times.


Figure 4. Time of sampling (in red) in the tidal cycle at Poughkeepsie.

### 2.5. Storm Water PCDD/Fs and PCBs.

Storm water samples were taken at Bayonne of $6 / 10 / 19$ and 10/16/19. It was starting to rain at Newark Airport when the $6 / 10$ sample was taken and raining rather hard when the $10 / 16$ sampling was being done.


Figure 21. Relative abundances of total PCDD/F TEQ at Bayonne storm water.


Figure 22. Relative abundances of PCB homologs at Bayonne storm water.

The 7/23/19 Belleville storm water sample (Bell) was taken as an earlier rain storm was ending. The sample taken on 10/3/19 was during a rain storm at the Newark Airport.


Figure 23. Relative abundances of PCDD/F TEQ at Belleville storm water.


Figure 54. Relative abundances of PCB homologs at Belleville storm water.
Sampling at the Kearny storm water site (Kea) occurred during an early morning rain shower. A sample was also taken on 12/2/19 but lost in a lab accident.


Figure 25. Relative abundances of PCDD/F TEQ at Kearny Point storm water.

A lab accident lost any possibility of detecting the mono-chlorobiphenyls in Kea-07/23/19.


Figure 26. Relative abundances of PCB homologs at Kearny Point storm water.

The highest total TEQ seen in water samples occurred in one of the Keyport storm water samples. Weather data from Newark International Airport indicates that it was dry on 5/28/19 but rainy on $12 / 9 / 19$. Congener patterns on the two occasions were very similar but total TEQs were different.

The suspended sediment particles trapped by the filter differed in abundance but had the same source.


Figure 27. Relative abundances of PCDD/F TEQ in Keyport storm water.


Figure 28. Relative abundances of PCB homologs in Keyport storm water.

Light rain was falling at Newark Airport on both sampling occasions at New Milford (NM). Total TEQ was high, particularly from the 12/2/19 event. Congener patterns on the two occasions were very similar.


Figure 29. Relative abundances of PCDD/F TEQ at New Milford storm water.


Figure 30. Relative abundances of PCB homologs at New Milford storm water. 6

Total TEQs of most of the water column samples shown above were dominated by congener \#6, ( $1,2,3,4,6,7,8-H P C D D)$. This congener could be produced by dechlorination of OCDD, which is always the most abundant dioxin congener, in terms of raw concentration instead of TEF adjusted values, of the seven that are measured.

Dichlorobiphenyl (PCB homolog \# 2) was the dominant PCB homolog group only in the two New Milford storm water samples (NM). In 16 of the TOPS sampling events the major congener of the dichlorobiphenls was 3,3'-dichlorobiphenyl (IUPAC 11). Prior to the use of method 1668 for PCBs this congener was unknown in environmental work. Printing shops, which may have been using diarylide pigments (often contaminated with 3,3'-DiCB) were found immediately upstream from one of the storm sewer sampling points. A partial exception to the 1979 PCB ban for inadvertently manufactured mono- and dichlorobiphenyls was made to the Federal Toxic Substances Control Act in 1986. State regulations may differ.

## Chapter 3. Sediments

For the evaluation of sediments in navigation channels and off-channel areas, sediment core samples were collected from six locations in NY-NJ Harbor: Buttermilk Channel, Elizabeth Channel, Port Jersey, Port Newark, South Brother Island Channel and Ward Point Bend near the mouth of the Raritan River (Figure 2). Generally, for each station, three samples were collected in the navigation channel and three samples were collected from the adjacent off-channel sediments. To represent sediments that are typically considered in dredged material testing, navigation channel sediment core samples were sub-sampled from a surface layer ( $0-10 \mathrm{~cm}$ ) and a deeper layer ( $20-30 \mathrm{~cm}$ ). Since offchannel sediments were expected to accumulate at slower rates, off-channel sediment samples were sub-sampled from a surface layer ( $0-4 \mathrm{~cm}$ ) and a deeper layer ( $6-10 \mathrm{~cm}$ ). In total 68 sediment samples were collected from in-channel and off-channel areas across the harbor.

Each of the 68 sediment samples were analyzed for PCB congeners and $2,3,7,8$-chlorine substituted PCDD/Fs. Additional sediment parameters of total organic carbon, soot (black carbon), dissolved organic carbon and Berylium-7 concentrations were also measured. A subset of 20 sediment samples were analyzed using polyethylene passive samplers to determine porewater concentrations for PCBs, PCDD/Fs, and black carbon. 28-day bioaccumulation tests were performed on all 68 of the samples following the protocols ${ }^{1}$ outlined in the Region 2 Testing Manual (USACE/USEPA 2016). These data were used to assess the accuracy of the CARP I models projections, and as discussed below, in the refinement of the CARP II models.

### 3.1. Sediment Core Samples and Map

Sediment cores were taken from six sediment groups. In each group three sediment cores were taken from the in the navigational channel (IC) and three were from adjacent off channel (OC) locations (except Port Newark when only one OC core was taken). Cores were taken using a HAPS gravity corer. The cores were sectioned into an upper (Over) "active" layer and a lower (Under) "deep" layer. The active layer or Over core sections were $0-10 \mathrm{~cm}$ in IC samples and $0-4 \mathrm{~cm}$ in OC samples. The deep layer or Under samples were usually from $20-30 \mathrm{~cm}$ in the IC cores and usually 6-10 cm from the OC or off channel cores.

[^0]After each deployment the HAPS corer and associated sampling equipment were thoroughly brushed and rinsed with site water. Sediment samples were analyzed for PCBs, PCDD/Fs, TOC, black carbon, particle size and $\mathrm{Be}-7$. Sediment wet weight, bulk density/dry weight and grain size distribution was also measured. Each sample was also be used for the 28-day bioaccumulation (Tissue) tests.

Table 6. Sediment core samples.

| Sample group and date | Abbr. | Nav. Status | Latitude | Longitude |
| :---: | :---: | :---: | :---: | :---: |
| Buttermilk_Channel_2-4/16/2019 | BMC | in channel | 40.67351 | -74.02555 |
| Buttermilk_Channel_3-4/16/2019 | BMC | in channel | 40.67703 | -74.03010 |
| Buttermilk_Channel_6-4/16/2019 | BMC | in channel | 40.67951 | -74.03024 |
| Buttermilk_Channel_3-4/16/2019 | BMC | off channel | 40.68249 | -74.02851 |
| Buttermilk_Channel_5-4/16/2019 | BMC | off channel | 40.68724 | -74.00736 |
| Buttermilk_Channel_6-4/16/2019 | BMC | off channel | 40.68559 | -74.01854 |
| Elizabeth Channel_1-5/16/2019 | EC | in channel | 40.68508 | -74.15281 |
| Elizabeth Channel_2-5/16/2019 | EC | in channel | 40.67494 | -74.14019 |
| Elizabeth Channel_3-5/16/2019 | EC | in channel | 40.66175 | -74.14715 |
| Elizabeth Channel_4-5/16/2019 | EC | off channel | 40.66545 | -74.13439 |
| Elizabeth Channel_5-5/16/2019 | EC | off channel | 40.65767 | -74.14086 |
| Elizabeth Channel_6-5/16/2019 | EC | off channel | 40.67912 | -74.12786 |
| Port Jersey 1-5/7/2019 | PJ | in channel | 40.66730 | -74.07487 |
| Port Jersey 2-5/7/2019 | PJ | in channel | 40.66711 | -74.07180 |
| Port Jersey 3-5/7/2019 | PJ | in channel | 40.66508 | -74.06992 |
| Port Jersey 4-5/7/2019 | PJ | off channel | 40.66327 | -74.06791 |
| Port Jersey 5-5/7/2019 | PJ | off channel | 40.66567 | -74.06331 |
| Port Jersey 6-5/7/2019 | PJ | off channel | 40.66957 | -74.08132 |
| Port_Newark_7-6/13/2019 | PN | in channel | 40.69254 | -74.13662 |
| Port_Newark_8-6/13/2019 | PN | in channel | 40.69505 | -74.13930 |
| Port_Newark_9-6/13/2019 | PN | in channel | 40.69798 | -74.15093 |
| Port_Newark_10-6/13/2019 | PN | off channel | 40.68388 | -74.12646 |
| Wards_Point_Bend_1-3/26/2019 | RB | in channel | 40.48556 | -74.24775 |
| Wards_Point_Bend_2-3/26/2019 | RB | in channel | 40.48433 | -74.25554 |
| Wards_Point_Bend_3-3/26/2019 | RB | in channel | 40.48445 | -74.26086 |
| Wards_Point_Bend_4-3/26/2019 | RB | off channel | 40.48604 | -74.24858 |
| Wards_Point_Bend_5-3/25/2019 | RB | off channel | 40.48412 | -74.24726 |
| Wards_Point_Bend_6-3/25/2019 | RB | off channel | 40.48539 | -74.25251 |
| South Brother_1-4/23/2019 | SB | in channel | 40.79239 | -73.89477 |
| South Brother_2-4/23/2019 | SB | in channel | 40.79407 | -73.89387 |
| South Brother_3-4/23/2019 | SB | in channel | 40.79667 | -73.89339 |
| South Brother_4-4/23/2019 | SB | off channel | 40.78859 | -73.89238 |
| South Brother_5-4/23/2019 | SB | off channel | 40.78645 | -73.89052 |
| South Brother_6-4/23/2019 | SB | off channel | 40.78979 | -73.89582 |



Figure 31. Sediment core sampling locations.

### 3.2. Sediment Grab Samples and Map.

Surficial sediment samples were collected throughout the harbor to evaluate how contaminant concentrations and sediment properties vary across spatial gradients. At 42 locations, surficial sediment samples were collected and analyzed for grain size, total organic carbon (TOC) and soot (black) carbon. Eight of these were to be analyzed for PCBs and PCDD/Fs but one the PCB extracts was lost in a laboratory accident.

Table 7. Sediment grab samples.

| Sample name | date | Abbr. | Nav. Status | Latitude | Longitude |
| :--- | ---: | :--- | :--- | :--- | :--- |
| Arthur Kill_1 | $9 / 20 / 2021$ | AK1 | in channel | 40.64151 | -74.19033 |
| Arthur Kill_3 | $10 / 21 / 2021$ | AK3 | in channel | 40.60400 | -74.20240 |
| Arthur Kill_4 | $10 / 21 / 2021$ | AK4 | in channel | 40.57830 | -74.20950 |
| Arthur Kill_5 | $10 / 21 / 2021$ | AK5 | in channel | 40.53040 | -74.24880 |
| East River_1 | $10 / 25 / 2021$ | ER1 | off channel | 40.79520 | -73.90290 |
| East River_2 | $10 / 25 / 2021$ | ER2 | off channel | 40.73340 | -73.96290 |
| East River_3 | $10 / 25 / 2021$ | ER3 | off channel | 40.69880 | -73.99940 |
| Harlem River_1 | $9 / 16 / 2021$ | HAR1 | off channel | 40.83899 | -73.93155 |


| Harlem River_2 | $9 / 16 / 2021$ | HAR2 | off channel | 40.79116 | -73.93196 |
| :--- | ---: | :--- | :--- | :--- | :--- |
| Hudson River_1 | $9 / 16 / 2021$ | HR1 | off channel | 40.94351 | -73.90973 |
| Hudson River_2 | $9 / 16 / 2021$ | HR2 | off channel | 41.19372 | -73.92225 |
| Hudson River_3 | $9 / 16 / 2021$ | HR3 | off channel | 41.05609 | -73.88911 |
| Hudson River_4 | $9 / 16 / 2021$ | HR4 | off channel | 40.99806 | -73.88597 |
| Hudson River_5 | $9 / 22 / 2021$ | HR5 | off channel | 40.82116 | -73.96830 |
| Hudson River_6 | $9 / 22 / 2021$ | HR6 | off channel | 40.76615 | -74.00509 |
| Hudson River_7 | $9 / 22 / 2021$ | HR7 | off channel | 40.71045 | -74.02617 |
| Jamaica Bay_1 | $10 / 21 / 2021$ | JB1 | off channel | 40.57160 | -73.94510 |
| Jamaica Bay_2 | $10 / 21 / 2021$ | JB2 | off channel | 40.62118 | -73.77878 |
| Jamaica Bay_3 | $10 / 21 / 2021$ | JB3 | off channel | 40.62650 | -73.86030 |
| Jamaica Bay_4 | $10 / 21 / 2021$ | JB4 | off channel | 40.59260 | -73.86330 |
| Kill Van Kull_1 | $9 / 20 / 2021$ | KVK1 | in channel | 40.64405 | -74.13007 |
| Kill Van Kull_2 | $9 / 20 / 2021$ | KVK2 | in channel | 40.65010 | -74.08208 |
| Long Island Sound_1 | $10 / 25 / 2021$ | LIS1 | off channel | 40.88944 | -73.72111 |
| Long Island Sound_2 | $10 / 25 / 2021$ | LIS2 | off channel | 40.80399 | -73.78640 |
| Long Island Sound_3 | $10 / 25 / 2021$ | LIS3 | off channel | 40.79620 | -73.85020 |
| Lower_Bay_1 | $10 / 21 / 2021$ | LB1 | off channel | 40.59310 | -74.02110 |
| Lower_Bay_2 | $10 / 21 / 2021$ | LB2 | off channel | 40.58570 | -74.05760 |
| Newark Bay_1 | $9 / 20 / 2021$ | NB1 | in channel | 40.71319 | -74.11184 |
| Newark Bay_2 | $9 / 20 / 2021$ | NB2 | off channel | 40.65538 | -74.15426 |
| Newark Bay_3 | $9 / 20 / 2021$ | NB3 | off channel | 40.68920 | -74.12087 |
| Upper Bay_1 | $9 / 20 / 2021$ | UB1 | off channel | 40.66169 | -74.02973 |
| Upper Bay_2 | $9 / 20 / 2021$ | UB2 | off channel | 40.65369 | -74.03019 |
| Upper Bay_3 | $9 / 22 / 2021$ | UB3 | off channel | 40.69438 | -74.03762 |
| Upper Bay_4 | $9 / 22 / 2021$ | UB4 | in channel | 40.67620 | -74.05347 |
| Upper Bay_5 | $9 / 22 / 2021$ | UB5 | off channel | 40.63107 | -74.06226 |
| Upper Bay_6 | $9 / 22 / 2021$ | UB6 | off channel | 40.61381 | -74.04735 |



Figure 32. Sediment grab samples.

### 3.3. Sediment Characterization

Sediment cores were characterized by grain size analysis, soot carbon (or black carbon), solids, and total organic carbon (TOC).

Table 8. Sediment characterization.


| EC-IC-Under | 48.1 | 14.9 | 31.4 | 6.2 | 1.4 | 0.1 | 42 | 2.6 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| EC-OC-Over | 53.4 | 15.7 | 26.7 | 4.1 | 0.8 | 0.1 | 43 | 2.3 |
| EC-OC-Under | 52.9 | 15.2 | 26.6 | 6.5 | 0.6 | 0.2 | 44.3 | 2.2 |
| PJ | 27.3 | 12.7 | 21.5 | 26.7 | 11.8 | 0.2 | 39.6 | 2.7 |
| PJ-IC-Over | 29.6 | 7.1 | 23.4 | 29.9 | 9.9 | 0.1 | 31.9 | 3.1 |
| PJ-IC-Under | 21.7 | 4.9 | 17.3 | 32.8 | 23.3 | 0.1 | 35.6 | 3.1 |
| PJ-OC-Over | 30.2 | 21.6 | 23.3 | 17.9 | 7.1 | 0.4 | 46.4 | 2.2 |
| PJ-OC-Under | 27.6 | 17.3 | 21.8 | 26.3 | 7 | 0.1 | 44.5 | 2.4 |
| PN | 45.8 | 25.1 | 21.4 | 7.6 | 0.5 | 0.2 | 45.5 | 2.5 |
| PN-IC-Over | 52.3 | 9.9 | 27.8 | 9.9 | 0.3 | 0.1 | 30.7 | 3.5 |
| PN-IC-Under | 46.9 | 11.6 | 28.6 | 12.5 | 0.7 | 0.1 | 36.2 | 3.1 |
| PN-OC-Over | 40.9 | 35.8 | 19.2 | 4.1 | ND | 0.3 | 54.4 | 1.7 |
| PN-OC-Under | 43.2 | 42.9 | 9.9 | 4 | ND | 0.3 | 60.8 | 1.5 |
| RB | 55.5 | 13.0 | 14.5 | 8.5 | 9.8 | 0.1 | 36.7 | 3.4 |
| RB-IC-Over | 73.1 | 5.3 | 8.3 | 8 | 5.4 | 0.1 | 30.9 | 4 |
| RB-IC-Under | 46.5 | 6 | 13.6 | 14.6 | 19.2 | 0.1 | 34.2 | 3.8 |
| RB-OC-Over | 47.4 | 23.9 | 19.8 | 5.1 | 5.8 | ND | 41.7 | 2.7 |
| RB-OC-Under | 55.1 | 16.7 | 16.4 | 6.1 | 8.6 | 0.1 | 39.8 | 3 |
| SB | 32.8 | 32.0 | 30.3 | 4.3 | 1.3 | 0.8 | 40.4 | 3.8 |
| SB-IC-Over | 41.8 | 14.2 | 37.1 | 4.8 | 2.1 | 0.1 | 24 | 4.5 |
| SB-IC-Under | 38.4 | 15.2 | 39.2 | 6.8 | 1 | 0.1 | 31.3 | 4 |
| SB-OC-Over | 26.1 | 48.5 | 22.7 | 2.3 | 1.4 | 1.2 | 51.8 | 3.2 |
| SB-OC-Under | 24.8 | 50.2 | 22 | 3.3 | 0.6 | 1.6 | 54.5 | 3.5 |

Beryllium-7 (in $\mathrm{pCi} / \mathrm{g}$ ) assesses the integrity of the cores. This short-lived isotope should have higher concentrations in the Over segment than in the Under ones. This condition was met, on average, for all samples except for the Raritan Bay (RB) Off-Channel samples where two of the three cores showed higher $\mathrm{Be}-7$ in the Under segments. Be-7 error bars are very broad relative to the observed concentrations so this parameter should be regarded cautiously.

Table 9. Beryllium-7.

|  | $\mathrm{Be}-7$ | $+/-$ |  | $\mathrm{Be}-7$ | $+/-$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
| BMC |  |  | PN |  |  |
| BMC-IC-Over | 1.907 | 0.977 | PN-IC-Over | 2.929 | 1.477 |
| BMC-IC-Under | 0.298 | 1.05 | PN-IC-Under | 0.128 | 1.086 |
| BMC-OC-Over | 1.784 | 1.232 | PN-OC-Over | 0.548 | 0.845 |
| BMC-OC-Under | 0 | 1.078 | PN-OC-Under | 0 | 0.326 |
| EC |  |  | RB |  |  |
| EC-IC-Over | 2.226 | 1.544 | RB-IC-Over | 5.342 | 1.791 |
| EC-IC-Under | 0.811 | 1.14 | RB-IC-Under | 0.345 | 1.578 |
| EC-OC-Over | 2.433 | 1.479 | RB-OC-Over | 0.238 | 1.183 |


| EC-OC-Under | 0.336 | 0.828 | RB-OC-Under | 0.289 | 1.551 |
| :--- | :---: | :---: | :---: | :---: | :---: |
| PJ |  |  |  |  | SB |
|  |  |  |  |  |  |
| PJ-IC-Over | 5.033 | 1.554 | SB-IC-Over | 7.991 | 2.369 |
| PJ-IC-Under | 1.393 | 1.206 | SB-IC-Under | 1.158 | 1.643 |
| PJ-OC-Over | 1.736 | 0.871 | SB-OC-Over | 1.32 | 1.184 |
| PJ-OC-Under | 0.81 | 0.831 | SB-OC-Under | 0.405 | 1.352 |

Sample group grain size proportions, averages of 2019 samples by sediment group.






Figure 33. Average grain size characteristics from the six clusters of sediment core.

Sediment characteristics of the 42 grab samples taken in 2021 are shown in Table 10.

Table 10. Grab sample sediment characteristics.

| Abbr | harbor_region | Grain size, \% |  |  |  |  | \% |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Fines | Fine Sand | Medium Sand | Coarse Sand | Gravel | Soot | Solids | TOC |
| AK1 | Upper Bay | 48.8 | 15.9 | 10.4 | 9 | 15.9 | 0.23 | 31.3 | 3.2 |
| AK3 | Arthur Kill | 6.5 | 4.1 | 27.9 | 20.4 | 41.1 | 0.15 | 76 | 0.7 |


| AK4 | Arthur Kill | 17.1 | 28.6 | 38.2 | 6.6 | 9.5 | 0.07 | 71.1 | 1.8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AK5 | Arthur Kill | 53.4 | 9.5 | 23.9 | 13.2 |  | 0.13 | 36.1 | 3.9 |
| ER1 | East River | 12.1 | 14.2 | 35.9 | 11.1 | 26.7 |  | 75.2 | 1.9 |
| ER2 | East River | 56.4 | 13.1 | 21.2 | 9.1 | 0.2 | 2.21 | 38.9 | 6 |
| ER3 | East River | 49.3 | 4.5 | 20.8 | 21.9 | 3.5 | 0.1 | 31.2 | 3.5 |
| HAR1 | Hudson River | 2.3 | 12.2 | 84.4 | 1.1 |  | 0.02 | 75.4 | 0.2 |
| HAR2 | Harlem River | 43 | 12 | 17.3 | 21.5 | 6.2 | 1.76 | 43.2 | 5.6 |
| HR1 | Hudson River | 22.2 | 50.9 | 24 | 1.2 | 1.7 | 0.04 | 68.2 | 0.6 |
| HR2 | Hudson River | 44 | 13.6 | 11.6 | 7.4 | 23.4 | 0.16 | 56.2 | 1.9 |
| HR3 | Hudson River | 66.6 | 8 | 17.3 | 7.9 | 0.2 | 0.17 | 42.1 | 2.2 |
| HR4 | Hudson River | 74.2 | 12 | 9 | 3.5 | 1.3 | 0.18 | 51 | 2 |
| HR5 | Hudson River | 47.9 | 4.5 | 13.5 | 26.3 | 7.8 | 0.12 | 36.6 | 2.4 |
| HR6 | Hudson River | 52.2 | 5.6 | 16.5 | 21.4 | 4.3 | 0.08 | 41.9 | 2.6 |
| HR7 | Hudson River | 33.8 | 32.6 | 18 | 14.6 | 1 | 0.11 | 34 | 2.5 |
| JB1 | Jamaica Bay | 26 | 58.9 | 5.2 | 3.1 | 6.8 | 0.02 | 59.9 | 1.4 |
| JB2 | Jamaica Bay | 45.1 | 5.1 | 16 | 23.3 | 10.5 | 0.1 | 20.2 | 5.9 |
| JB3 | Jamaica Bay | 15.1 | 83.8 | 0.3 | 0.4 | 0.4 | 0.03 | 72.8 | 0.3 |
| JB4 | Jamaica Bay | 53.7 | 29.3 | 12.5 | 4.5 |  | 0.1 | 42.6 | 2.8 |
| KVK1 | Kill Van Kull | 14.1 | 48.1 | 34.3 | 2.8 | 0.7 | 0.03 | 58.8 | 0.9 |
| KVK2 | Kill Van Kull | 44.75 | 47.95 | 5.65 | 1.65 |  | 0.07 | 54.5 | 1 |
| LB1 | Lower Bay | 58.6 | 35.9 | 4.9 | 0.6 |  | 0.07 | 52.5 | 1.5 |
| LB2 | Lower Bay | 24.5 | 63.6 | 7.2 | 4.1 | 0.6 | 0.02 | 64.8 | 0.8 |
| LIS1 | Long Island Sound | 57.7 | 6.5 | 21.7 | 13.2 | 0.9 | 0.16 | 34.7 | 3.2 |
| LIS2 | Long Island Sound | 19.1 | 53.5 | 16.2 | 5.4 | 5.8 | 0.22 | 71.1 | 1.2 |
| LIS3 | Long Island Sound | 15 | 46.6 | 24.7 | 7.2 | 6.5 | 0.03 | 80.4 | 0.4 |
| NB1 | Arthur Kill | 21.4 | 66.5 | 9 | 1.6 | 1.5 | 0.31 | 60.2 | 2.2 |
| NB2 | Newark Bay | 32.9 | 62.6 | 3.7 | 0.8 |  | 0.47 | 62.8 | 0.8 |
| NB3 | Newark Bay | 39 | 23.8 | 22.7 | 10.6 | 3.9 | 0.18 | 42.4 | 2.5 |
| RB1 | Raritan Bay | 30.5 | 21.7 | 32.4 | 14.7 | 0.7 | 0.1 | 41 | 3.4 |
| RB2 | Raritan Bay | 46.6 | 13.9 | 21.6 | 15.2 | 2.7 | 0.14 | 45.3 | 2.5 |
| RB3 | Raritan Bay | 0.8 | 90.8 | 8 | 0.4 |  | 0.03 | 76.7 | 0.1 |
| RB4 | Raritan Bay | 8.2 | 89.6 | 0.5 | 0.7 | 1 |  | 71.6 | 0.2 |
| RB5 | Raritan Bay | 43.45 | 53.45 | 2.5 | 0.6 |  | 0.07 | 64.1 | 0.8 |
| RR1 | Raritan River | 30.1 | 4.9 | 41.4 | 22.7 | 0.9 | 0.14 | 31.3 | 4.2 |
| UB1 | Upper Bay | 29.9 | 43.5 | 16.2 | 8.6 | 1.8 | 1.14 | 51.5 | 3.2 |
| UB2 | Newark Bay | 75.1 | 12 | 9.4 | 3.5 |  | 0.22 | 46.9 | 2 |
| UB3 | Upper Bay | 16.6 | 11.2 | 44.6 | 13.6 | 14 | 0.09 | 49 | 1.7 |
| UB4 | Upper Bay | 75 | 10.1 | 12.4 | 2.5 |  | 0.1 | 38 | 2.6 |
| UB5 | Upper Bay | 6.9 | 89.1 | 3.2 | 0.5 | 0.3 |  | 73.8 | 0.2 |
| UB6 | Harlem River | 10.6 | 61.3 | 13.9 | 5.5 | 8.7 | 0.19 | 74 | 0.2 |

The question of whether navigational versus non-navigational channels (In Channel v Off Channel) sites can be seen in sediment characteristic data was addressed by use of principal component analysis calculated in the R statistics package. Data from 2019 and 2021 samples are used.

Eight principal components were calculated. The first four accounted for $90 \%$ of the total variance. The primary axis, accounting for $39.5 \%$ of the total variability, represents TOC, fines, coarse sand and the absence of solids. The second principal component ( $21 \%$ of the total variance) represents gravel, coarse sand, and the absence of fines. Red values are negative.

Table 11. Principal components of sediment characteristics, from grab and core samples.

|  | PC1 | PC2 | PC3 | PC4 | PC5 | PC6 | PC7 | PC8 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Standard deviation | 1.7777 | 1.2964 | 1.1015 | 1.0869 | 0.6520 | 0.5258 | 0.2502 | 0.0020 |
| Proportion of Variance | 0.3951 | 0.2101 | 0.1517 | 0.1477 | 0.0531 | 0.0346 | 0.0078 | 0 |
| Cumulative Proportion | 0.3951 | 0.6051 | 0.7568 | 0.9045 | 0.9576 | 0.9922 | 1 | 1 |


|  | PC1 | PC2 | PC3 | PC4 | PC5 | PC6 | PC7 | PC8 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| coarse sand | 0.3175 | 0.5087 | 0.2158 | 0.0362 | 0.5611 | 0.4398 | 0.1282 | 0.2604 |
| fine sand | 0.5166 | 0.1355 | 0.1720 | 0.0715 | 0.3315 | 0.3654 | 0.0309 | 0.6601 |
| fines | 0.3657 | 0.4206 | 0.1849 | 0.3758 | 0.3101 | 0.3322 | 0.0682 | 0.5503 |
| gravel | 0.1321 | 0.6268 | 0.2207 | 0.2351 | 0.5441 | 0.3458 | 0.1404 | 0.2244 |
| medium sand | 0.0721 | 0.1262 | 0.3105 | 0.8348 | 0.1909 | 0.0569 | 0.0502 | 0.3787 |
| solids | 0.5171 | 0.1673 | 0.0420 | 0.0183 | 0.3106 | 0.3515 | 0.6946 | 0.0001 |
| soot | 0.0679 | 0.2457 | 0.7891 | 0.2449 | 0.2209 | 0.3358 | 0.3013 | 0.0000 |
| TOC | 0.4515 | 0.2210 | 0.3465 | 0.1996 | 0.0366 | 0.4512 | 0.6184 | 0.0002 |

The shaded circles represent 95\% confidence intervals for In Channel (blue) and Off Channel (gold) observations. The extent of overlap suggests that the two kinds of samples are not statistically different with respect to sediment quality.


Figure 34. Principal component analysis of sediment characteristics from In-and Off-Channel samples.

### 3.4. PCDD/Fs and PCBs in Sediment Cores.

Sediment cores were collected from six areas around the harbor. Sampling cores were usually taken at three sites within navigational channels (IC, In-Channel) and three taken out of navigational channels (OC, Off-Channel). Each of the cores was divided into in an upper (Over) portion and a lower (Under) portion.

Since there was one primary source of most of the PCDD/F TEQ, the former Diamond Alkali plant on the Passaic River, there was more of a spatial gradient in total TEQ across the sediment groups than there was for total PCBs where there were multiple sources.

Average and median concentrations of PCDD/Fs in sediment and tissue were higher in Off-Channel samples than from the In-Channel ones. PCBs were higher in Off-Channel tissues but not in OffChannel sediment. Under samples were higher in PCDD/Fs and PCBs for sediment and for tissue.


Figure 35. Graphical representation of PCDD/Fs in sediments, pg/g total TEQ. Values are averages of the triplicate replicates from each site.


Figure 36. Graphical representation of PCBs in sediments, pg/g. Values are averages of the triplicate replicates from each site.

### 3.5. PCDD/Fs in Sediment Cores

The following figures show relative abundances of TEQ PCDD/Fs (in pg/g dry weight) from the six sediment groups; Buttermilk Channel (BMC), Elizabeth Channel (EC), Port Jersey (PJ), Port Newark (PN), Raritan River (RR), and South Brother (SB). Sampling at each sediment group consisted of three cores taken in a dredged area (IC or In Channel), and three taken off channel (OC). The cores were sectioned into upper (Over) and a lower (Under) portions. In the figures the TEQs were averaged.

As expected, the Newark Bay sediment groups (EC and PN) had the highest concentrations and the highest proportional contribution by 2,3,7,8-TCDD. The relative abundances of all the congeners in the two Newark Bay groups were virtually identical. Port Newark and South Brother had higher average total TEQs in the in channel (IC) samples but off channel averages (OC) were higher in the others. In every case the average total TEQs were higher in the lower portions
(Under) of the cores than in the upper parts (Over). In every sediment group 2,3,7,8-TCDD (\# 1) and 2,3,7,8-TCDF (\#8) were the greatest contributors to TEQ. Congener \#6 (1,2,3,4,6,7,8-HPCDD) which was usually the single greatest contributor to TEQ in the water samples was usually the third greatest contributor in the sediment samples.

Table 12. PDCC/Fs in sediment. Blank corrected TEQ, pg/g. Values are averages of the replicates from each site. See Table 3 for Order numbers 1-17.

| Sed. Group | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | sum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BMC | 4.23 | 1.62 | 0.21 | 0.85 | 0.70 | 1.67 | 0.51 | 2.21 | 0.10 | 1.23 | 0.35 | 0.56 | 0.03 | 0.25 | 0.48 | 0.03 | 0.02 | 15.0 |
| BMC-IC-Over | 2.26 | 1.34 | 0.19 | 0.71 | 0.57 | 1.41 | 0.46 | 1.51 | 0.07 | 1.09 | 0.27 | 0.42 | 0.02 | 0.20 | 0.39 | 0.02 | 0.01 | 11.0 |
| BMC-IC-Under | 2.97 | 1.49 | 0.19 | 0.80 | 0.69 | 2.07 | 0.60 | 1.78 | 0.09 | 1.16 | 0.32 | 0.56 | 0.02 | 0.23 | 0.45 | 0.03 | 0.02 | 13.5 |
| BMC-OC-Over | 2.51 | 1.33 | 0.15 | 0.71 | 0.57 | 1.30 | 0.41 | 1.75 | 0.09 | 1.08 | 0.28 | 0.45 | 0.03 | 0.20 | 0.40 | 0.02 | 0.02 | 11.3 |
| BMC-OC-Under | 9.16 | 2.32 | 0.29 | 1.18 | 0.97 | 1.90 | 0.55 | 3.78 | 0.14 | 1.59 | 0.51 | 0.81 | 0.04 | 0.38 | 0.69 | 0.04 | 0.03 | 24.4 |
| EC | 30.75 | 2.84 | 0.36 | 1.31 | 1.09 | 2.62 | 0.85 | 4.32 | 0.17 | 2.61 | 1.02 | 3.43 | 0.04 | 0.56 | 1.79 | 0.07 | 0.08 | 53.9 |
| EC-IC-Over | 19.80 | 2.58 | 0.34 | 1.32 | 1.09 | 2.69 | 0.89 | 3.51 | 0.15 | 2.24 | 0.83 | 2.24 | 0.03 | 0.48 | 1.43 | 0.06 | 0.06 | 39.7 |
| EC-IC-Under | 28.12 | 2.75 | 0.37 | 1.32 | 1.12 | 2.83 | 0.86 | 4.02 | 0.16 | 2.48 | 0.97 | 3.23 | 0.05 | 0.55 | 1.68 | 0.07 | 0.07 | 50.7 |
| EC-OC-Over | 39.83 | 3.25 | 0.37 | 1.31 | 1.06 | 2.45 | 0.87 | 4.98 | 0.19 | 2.93 | 1.15 | 4.14 | 0.05 | 0.59 | 1.95 | 0.08 | 0.08 | 65.3 |
| EC-OC-Under | 37.00 | 2.86 | 0.34 | 1.29 | 1.08 | 2.35 | 0.78 | 4.97 | 0.18 | 2.87 | 1.17 | 4.26 | 0.05 | 0.62 | 2.18 | 0.08 | 0.09 | 62.2 |
| PJ | 4.41 | 2.02 | 0.26 | 1.06 | 0.85 | 1.95 | 0.60 | 2.44 | 0.11 | 1.58 | 0.39 | 0.64 | 0.03 | 0.29 | 0.57 | 0.03 | 0.03 | 17.3 |
| PJ-IC-Over | 3.05 | 1.98 | 0.29 | 1.13 | 0.87 | 2.15 | 0.66 | 2.17 | 0.10 | 1.77 | 0.39 | 0.56 | 0.02 | 0.29 | 0.57 | 0.03 | 0.04 | 16.1 |
| PJ-IC-Under | 3.86 | 2.33 | 0.29 | 1.19 | 0.98 | 2.21 | 0.71 | 2.46 | 0.11 | 1.66 | 0.44 | 0.69 | 0.03 | 0.31 | 0.64 | 0.03 | 0.03 | 18.0 |
| PJ-OC-Over | 4.81 | 1.74 | 0.24 | 0.90 | 0.76 | 1.59 | 0.48 | 2.39 | 0.09 | 1.38 | 0.35 | 0.57 | 0.02 | 0.25 | 0.49 | 0.03 | 0.02 | 16.1 |
| PJ-OC-Under | 5.55 | 2.03 | 0.25 | 1.05 | 0.82 | 1.87 | 0.57 | 2.66 | 0.11 | 1.52 | 0.40 | 0.71 | 0.03 | 0.31 | 0.57 | 0.03 | 0.03 | 18.5 |
| PN | 48.12 | 2.99 | 0.35 | 1.37 | 1.10 | 2.52 | 0.83 | 5.44 | 0.18 | 2.98 | 1.32 | 4.81 | 0.05 | 0.67 | 2.41 | 0.08 | 0.13 | 75.4 |
| PN-IC-Over | 41.57 | 3.43 | 0.45 | 1.61 | 1.40 | 3.19 | 1.04 | 4.93 | 0.19 | 3.04 | 1.38 | 4.37 | 0.06 | 0.74 | 2.55 | 0.09 | 0.10 | 70.1 |
| PN-IC-Under | 45.03 | 3.40 | 0.42 | 1.66 | 1.32 | 3.07 | 1.01 | 5.66 | 0.22 | 3.36 | 1.44 | 5.14 | 0.06 | 0.77 | 2.56 | 0.09 | 0.18 | 75.4 |
| PN-OC-Over | 49.30 | 2.18 | 0.22 | 0.82 | 0.68 | 1.41 | 0.49 | 4.94 | 0.14 | 2.35 | 1.06 | 4.35 | 0.04 | 0.47 | 2.00 | 0.07 | 0.10 | 70.6 |
| PN-OC-Under | 62.00 | 2.12 | 0.18 | 0.82 | 0.55 | 1.25 | 0.42 | 6.13 | 0.14 | 2.63 | 1.18 | 5.20 | 0.04 | 0.51 | 2.18 | 0.07 | 0.13 | 85.5 |
| RB | 9.06 | 3.72 | 0.43 | 2.06 | 1.62 | 2.89 | 1.70 | 3.83 | 0.25 | 2.60 | 0.87 | 1.65 | 0.07 | 0.65 | 1.13 | 0.07 | 0.05 | 32.6 |
| RB-IC-Over | 7.17 | 3.28 | 0.41 | 1.52 | 1.35 | 2.70 | 1.67 | 3.04 | 0.21 | 2.27 | 0.70 | 1.32 | 0.06 | 0.54 | 0.97 | 0.06 | 0.04 | 27.3 |
| RB-IC-Under | 10.06 | 3.67 | 0.46 | 1.83 | 1.56 | 3.12 | 1.91 | 3.61 | 0.27 | 2.68 | 0.86 | 1.59 | 0.06 | 0.64 | 1.11 | 0.07 | 0.05 | 33.5 |
| RB-OC-Over | 6.96 | 3.07 | 0.35 | 1.55 | 1.29 | 2.42 | 1.40 | 2.98 | 0.21 | 2.27 | 0.74 | 1.39 | 0.06 | 0.54 | 0.93 | 0.06 | 0.04 | 26.2 |
| RB-OC-Under | 11.71 | 4.88 | 0.49 | 3.42 | 2.31 | 3.26 | 1.76 | 5.74 | 0.31 | 3.17 | 1.17 | 2.31 | 0.10 | 0.90 | 1.51 | 0.10 | 0.07 | 43.2 |
| SB | 3.00 | 2.51 | 0.29 | 1.02 | 0.95 | 1.84 | 0.60 | 2.72 | 0.16 | 1.89 | 0.56 | 0.78 | 0.03 | 0.44 | 0.76 | 0.03 | 0.03 | 17.6 |
| SB-IC-Over | 2.73 | 2.55 | 0.35 | 1.25 | 1.13 | 2.41 | 0.78 | 2.51 | 0.14 | 1.80 | 0.53 | 0.72 | 0.03 | 0.40 | 0.75 | 0.04 | 0.03 | 18.2 |
| SB-IC-Under | 3.67 | 3.14 | 0.36 | 1.33 | 1.20 | 2.60 | 0.87 | 3.15 | 0.20 | 2.29 | 0.68 | 0.94 | 0.04 | 0.55 | 0.93 | 0.04 | 0.04 | 22.0 |
| SB-OC-Over | 2.26 | 1.87 | 0.20 | 0.69 | 0.63 | 1.26 | 0.41 | 2.19 | 0.14 | 1.48 | 0.41 | 0.63 | 0.03 | 0.32 | 0.53 | 0.03 | 0.02 | 13.1 |
| SB-OC-Under | 3.25 | 2.50 | 0.27 | 0.87 | 0.87 | 1.29 | 0.40 | 2.95 | 0.16 | 1.96 | 0.61 | 0.80 | 0.03 | 0.47 | 0.81 | 0.03 | 0.02 | 17.3 |



Figure 37. Average PCDD/F TEQ relative abundances from six sediment group.


Figure 38. Sediment BMC PCDD/F TEQ relative abundances.


Figure 39. Sediment EC PCDD/F TEQ relative abundances.


Figure 40. Sediment PJ PCDD/F TEQ relative abundances.


Figure 41. Sediment PN PCDD/F TEQ relative abundances.


Figure 42. Sediment RB PCDD/F TEQ relative abundances.


Figure 43. Sediment SB PCDD/F TEQ relative abundances.

### 3.6. PCBs in Sediment Cores.

Table 13. PCBs in Sediments, ng/g averages. Values are averages of the triplicate) replicates from each site.

| Sed. Group | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | sum |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BMC | 4.0 | 33.1 | 79.2 | 78.9 | 43.5 | 28.5 | 13.3 | 5.1 | 2.2 | 1.3 | 289.2 |
| BMC-IC-Over | 5.2 | 37.6 | 79.8 | 72.7 | 36.2 | 22.2 | 10.2 | 4.2 | 1.9 | 1.1 | 271.0 |
| BMC-IC-Under | 3.8 | 28.6 | 62.0 | 58.0 | 32.4 | 22.0 | 10.9 | 4.2 | 1.9 | 1.1 | 224.7 |
| BMC-OC-Over | 3.8 | 29.5 | 65.8 | 65.2 | 36.5 | 23.6 | 11.1 | 4.0 | 1.7 | 1.1 | 242.2 |
| BMC-OC-Under | 3.1 | 35.5 | 109.0 | 121.9 | 71.6 | 48.5 | 22.2 | 8.2 | 3.4 | 2.0 | 425.2 |
| EC | 4.2 | 35.1 | 81.1 | 98.4 | 71.2 | 52.9 | 23.0 | 10.0 | 4.7 | 2.3 | 382.9 |
| EC-IC-Over | 5.8 | 45.2 | 96.5 | 102.7 | 60.6 | 41.6 | 19.4 | 7.7 | 3.6 | 2.2 | 385.2 |
| EC-IC-Under | 4.0 | 35.4 | 83.9 | 99.8 | 66.0 | 46.8 | 22.4 | 8.5 | 3.3 | 2.3 | 372.4 |
| EC-OC-Over | 3.4 | 29.6 | 71.2 | 95.1 | 68.8 | 50.2 | 23.9 | 15.5 | 8.5 | 2.3 | 368.5 |
| EC-OC-Under | 3.4 | 30.2 | 71.9 | 95.5 | 90.9 | 75.1 | 26.8 | 8.8 | 3.8 | 2.4 | 408.9 |
| PJ | 6.9 | 53.4 | 109.8 | 99.0 | 49.3 | 31.9 | 14.9 | 5.9 | 2.4 | 1.5 | 375.0 |
| PJ-IC-Over | 9.0 | 65.6 | 127.8 | 105.3 | 48.4 | 28.1 | 12.3 | 5.8 | 2.6 | 1.5 | 406.3 |
| PJ-IC-Under | 8.3 | 61.9 | 122.2 | 105.7 | 50.8 | 32.5 | 14.7 | 6.4 | 2.9 | 1.7 | 407.1 |
| PJ-OC-Over | 5.0 | 41.1 | 87.8 | 84.9 | 44.2 | 30.6 | 14.5 | 5.0 | 1.9 | 1.2 | 316.3 |
| PJ-OC-Under | 6.0 | 48.5 | 104.7 | 99.8 | 51.9 | 34.5 | 16.7 | 6.3 | 2.4 | 1.5 | 372.3 |
| PN | 3.5 | 31.8 | 76.2 | 102.6 | 71.9 | 51.7 | 26.3 | 11.2 | 3.9 | 2.0 | 381.0 |
| PN-IC-Over | 4.9 | 42.2 | 99.0 | 122.0 | 79.0 | 56.1 | 27.9 | 10.5 | 3.7 | 2.3 | 447.8 |

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| PN-IC-Under | 4.1 | 36.4 | 84.4 | 113.8 | 81.5 | 61.0 | 32.4 | 15.2 | 5.5 | 2.6 | 436.8 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| PN-OC-Over | 1.6 | 16.2 | 41.2 | 59.8 | 44.5 | 30.6 | 15.5 | 6.3 | 2.1 | 1.2 | 218.9 |
| PN-OC-Under | 1.3 | 14.5 | 42.8 | 72.6 | 55.7 | 36.9 | 16.8 | 6.7 | 1.9 | 1.1 | 250.4 |
| RB | 2.7 | 22.8 | 58.6 | 80.8 | 60.7 | 40.4 | 18.5 | 7.4 | 3.8 | 2.8 | 298.5 |
| RB-IC-Over | 3.1 | 25.9 | 58.9 | 69.1 | 46.7 | 32.8 | 15.2 | 5.6 | 2.6 | 1.6 | 261.4 |
| RB-IC-Under | 2.9 | 23.1 | 55.4 | 69.5 | 53.0 | 37.0 | 17.1 | 6.2 | 2.8 | 1.8 | 268.9 |
| RB-OC-Over | 2.1 | 18.3 | 43.9 | 58.1 | 46.6 | 31.9 | 15.7 | 6.4 | 3.5 | 2.6 | 229.1 |
| RB-OC-Under | 2.5 | 23.9 | 77.4 | 130.2 | 99.0 | 61.1 | 26.6 | 11.7 | 6.6 | 5.6 | 444.7 |
| SB | 3.8 | 29.8 | 68.9 | 73.1 | 53.0 | 45.9 | 22.0 | 7.6 | 3.3 | 2.2 | 309.6 |
| SB-IC-Over | 6.9 | 52.1 | 113.5 | 103.5 | 51.6 | 48.1 | 24.1 | 7.3 | 3.0 | 1.9 | 412.1 |
| SB-IC-Under | 5.4 | 40.1 | 88.2 | 86.7 | 56.2 | 47.4 | 22.6 | 8.6 | 4.0 | 2.8 | 362.0 |
| SB-OC-Over | 2.1 | 16.7 | 39.3 | 46.7 | 47.9 | 41.1 | 16.8 | 5.9 | 2.6 | 1.9 | 220.9 |
| SB-OC-Under | 1.6 | 15.1 | 43.2 | 59.9 | 55.6 | 46.9 | 23.8 | 8.4 | 3.4 | 2.2 | 260.1 |



Figure 44. Average sediment PCB homolog relative abundances from six sediment groups.


Figure 45. Sediment BMC PCB homolog relative abundances.


Figure 46. Sediment EC PCB homolog relative abundances.


Figure 47. Sediment PJ PCB homolog relative abundances.


Figure 48. Sediment PN PCB homolog relative abundances.


Figure 49. Sediment RB PCB homolog relative abundances.


Figure 50. Sediment SB PCB homolog relative abundances.

## Chapter 4. Tissue



Sediment samples at the six harbor locations were collected for use in 28-day bioaccumulation tests using the dredged material test organism, Neanthes virens. At each site, three in-channel and three offchannel samples at each of four depth intervals were analyzed. In addition to measuring contaminant concentrations and lipid content of the worms at the end of the 28 -day exposure period, the weights of the worms were measured at the beginning and end of the 28 -day exposure period (to evaluate average growth rates). Sediment TOC at the beginning and end of the 28 -day period were also be measured and used in estimating the digestible organic carbon content of the sediment.

The bioaccumulation tests were comprised of the individual 84 test samples and 3 replicates of a control sediment. The methods for the bioaccumulation tests are a modification of the testing procedures outlined in the USACE/EPA Regional Testing Manual (USACE, 2017). Two important adaptions were made: 1) The volume of sediment per test organism was similar but the tests used 2.5 -gallon tanks instead of the standard 10 gallon tanks. This is needed to reduce the number of sediment samples needed to achieve the per sample volume of sediment required. Here we are analyzing small discrete sediment intervals of only a few centimeters instead of the typically large composite samples analyzed for HARS suitability testing. 2) There could be no replicates of the field samples.

The lab performing the bioassays, AquaSurvey, depurated and froze the specimens prior to shipping them to Axys for chemical analysis. The chemical lab, Axys, homogenized the samples prior to completing their analysis.

The potential for bioaccumulation of PCDD/Fs and PCBs from sediments was assessed by exposing Neanthes virens to sediments. Toxicity of sediments was assessed by recording survival and weight of Neanthes.

### 4.1. Lipids

Table 14. Evaluations of the body burden of aquatic organisms may be assisted by knowing their lipid composition.

|  | \% lipid |  | \% lipid |
| :---: | :---: | :---: | :---: |
| BMC | 1.41 | PN | 1.65 |
| BMC-IC-Over | 1.7 | PN-IC-Over | 1.73 |


| BMC-IC-Under | 1.2 | PN-IC-Under | 1.54 |
| :--- | :---: | :--- | :---: |
| BMC-OC-Over | 1.4 | PN-OC-Over | 1.66 |
| BMC-OC-Under | 1.35 | PN-OC-Under | 1.67 |
| EC | 2.16 | RB | 1.26 |
| EC-IC-Over | 2.47 | RB-IC-Over | 1.46 |
| EC-IC-Under | 1.6 | RB-IC-Under | 1.21 |
| EC-OC-Over | 2.11 | RB-OC-Over | 1.17 |
| EC-OC-Under | 2.44 | RB-OC-Under | 1.2 |
| PJ | 1.59 | SB | 1.74 |
| PJ-IC-Over | 1.48 | SB-IC-Over | 1.87 |
| PJ-IC-Under | 1.43 | SB-IC-Under | 1.2 |
| PJ-OC-Over | 2.04 | SB-OC-Over | 1.78 |
| PJ-OC-Under | 1.43 | SB-OC-Under | 2.12 |

### 4.2. Toxicity

The Newark Bay sites (PJ and EC) were perhaps a little less toxic than the samples from Raritan Bay (RB), South Brother Island (SB), and the Buttermilk Channel (BMC). Only one sample from the Elizabeth Channel (EC) was reported for toxicity.

Table 15. Toxicity.

| sample | \% <br> Survival | Tissue Mass (wet wt in g) |
| :--- | :---: | :---: |
| BMC-IC-Over | 92 | 31 |
| BMC-IC-Under | 92 | 33 |
| BMC-OC-Over | 79 | 31 |
| BMC-OC-Under | 92 | 28 |
| EC-IC-Under | 100 | 37 |
| PJ-IC-Over | 100 | 40 |
| PJ-IC-Under | 96 | 32 |
| PJ-OC-Over | 100 | 35 |
| PJ-OC-Under | 96 | 32 |
| PN-IC-Over | 100 | 38 |
| PN-IC-Under | 83 | 31 |
| PN-OC-Over | 100 | 35 |
| PN-OC-Under | 100 | 41 |
| RB-IC-Over | 96 | 39 |
| RB-IC-Under | 79 | 30 |
| RB-OC-Over | 79 | 26 |
| RB-OC-Under | 79 | 27 |
| SB-IC-Over | 90 | 36 |


| SB-IC-Under | 88 | 30 |
| :--- | :--- | :--- |
| SB-OC-Over | 90 | 33 |
| SB-OC-Under | 81 | 30 |

### 4.3 PCDD/Fs in Tissue

Data quality for PCDD/Fs in tissues was weaker than for PCDD/Fs in sediments.
Table 16. Averages of the triplicate replicates from each site, fg/g TEQ wet wt. Table 3 has the chemical names of the 1-17 congeners indicated.

| Sed. Group | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | sum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BMC | 146 | 122 | 3.2 | 17.2 | 9.0 | 10.9 | 1.5 | 104 | 4.6 | 54 | 6.5 | 9.8 | 2.5 | 5.6 | 4.5 | 0.1 | 0.1 | 502 |
| BMC-IC-Over | 95 | 123 | 5.7 | 16.5 | 8.2 | 12.2 | 1.8 | 86 | 3.8 | 52 | 5.9 | 9.1 | 2.2 | 4.0 | 5.1 | 0.2 | 0.1 | 430 |
| BMC-IC-Under | 98 | 80 | 0.0 | 13.7 | 7.6 | 8.1 | 1.2 | 86 | 4.1 | 44 | 5.4 | 8.6 | 6.1 | 3.3 | 3.3 | 0.0 | 0.1 | 369 |
| BMC-OC-Over | 127 | 152 | 5.4 | 21.4 | 9.2 | 13.5 | 1.6 | 110 | 4.8 | 62 | 6.5 | 9.7 | 0.0 | 9.4 | 6.0 | 0.2 | 0.2 | 539 |
| BMC-OC-Under | 263 | 133 | 1.8 | 17.1 | 10.8 | 9.8 | 1.5 | 135 | 5.8 | 59 | 8.3 | 12.0 | 1.8 | 5.7 | 3.7 | 0.0 | 0.1 | 668 |
| EC | 989 | 156 | 5.1 | 23.8 | 13.3 | 11.0 | 1.6 | 185 | 7.8 | 108 | 15.8 | 33.5 | 3.2 | 8.7 | 6.8 | 0.3 | 0.2 | 1569 |
| EC-IC-Over | 828 | 136 | 3.3 | 23.2 | 11.7 | 11.7 | 1.8 | 191 | 6.7 | 104 | 11.6 | 25.4 | 0.0 | 6.7 | 6.2 | 0.0 | 0.2 | 1368 |
| EC-IC-Under | 791 | 140 | 2.7 | 22.3 | 10.5 | 10.4 | 1.6 | 161 | 6.6 | 87 | 12.2 | 25.7 | 2.3 | 4.4 | 5.7 | 0.0 | 0.2 | 1283 |
| EC-OC-Over | 993 | 161 | 4.1 | 23.9 | 14.6 | 11.2 | 1.6 | 194 | 8.9 | 117 | 16.9 | 33.4 | 3.5 | 11.1 | 7.1 | 0.4 | 0.2 | 1602 |
| EC-OC-Under | 1343 | 186 | 10.3 | 25.8 | 16.3 | 10.6 | 1.6 | 194 | 9.2 | 126 | 22.4 | 49.5 | 7.2 | 12.7 | 8.4 | 0.8 | 0.2 | 2024 |
| PJ | 254 | 196 | 15.4 | 29.7 | 21.6 | 11.7 | 1.7 | 138 | 7.9 | 100 | 16.2 | 19.2 | 12.8 | 13.9 | 4.9 | 0.8 | 0.1 | 845 |
| PJ-IC-Over | 151 | 105 | 7.1 | 23.7 | 14.9 | 10.5 | 1.7 | 118 | 6.0 | 81 | 9.4 | 12.9 | 8.5 | 7.7 | 3.8 | 0.3 | 0.1 | 561 |
| PJ-IC-Under | 169 | 124 | 8.7 | 19.6 | 15.3 | 12.4 | 2.1 | 110 | 5.1 | 64 | 7.5 | 10.1 | 7.7 | 6.5 | 4.6 | 0.0 | 0.1 | 567 |
| PJ-OC-Over | 268 | 176 | 13.7 | 27.5 | 18.4 | 11.2 | 1.5 | 150 | 7.9 | 96 | 14.8 | 18.2 | 9.3 | 12.8 | 4.6 | 0.7 | 0.1 | 831 |
| PJ-OC-Under | 428 | 380 | 32.1 | 47.9 | 37.9 | 12.7 | 1.7 | 175 | 12.6 | 160 | 33.1 | 35.5 | 25.8 | 28.7 | 6.5 | 2.2 | 0.2 | 1419 |
| PN | 2126 | 167 | 5.6 | 20.2 | 11.8 | 11.5 | 1.7 | 256 | 9.1 | 118 | 18.7 | 50.5 | 1.0 | 7.7 | 9.6 | 0.3 | 0.3 | 2814 |
| PN-IC-Over | 1393 | 148 | 3.0 | 25.6 | 12.2 | 12.2 | 1.7 | 207 | 7.8 | 104 | 13.1 | 31.1 | 0.0 | 8.8 | 7.3 | 0.4 | 0.2 | 1977 |
| PN-IC-Under | 1430 | 138 | 0.0 | 16.4 | 10.7 | 9.1 | 1.4 | 208 | 8.4 | 99 | 15.6 | 37.3 | 4.0 | 8.7 | 6.3 | 0.2 | 0.2 | 1993 |
| PN-OC-Over | 1720 | 201 | 9.4 | 19.1 | 11.7 | 12.0 | 1.7 | 235 | 8.0 | 99 | 16.4 | 47.5 | 0.0 | 8.5 | 10.6 | 0.0 | 0.3 | 2401 |
| PN-OC-Under | 3960 | 180 | 10.2 | 19.6 | 12.4 | 12.6 | 2.1 | 372 | 12.0 | 168 | 29.8 | 86.2 | 0.0 | 4.8 | 14.2 | 0.7 | 0.4 | 4885 |
| RB | 233 | 116 | 3.1 | 19.6 | 12.1 | 15.1 | 3.9 | 129 | 6.3 | 68 | 8.7 | 15.1 | 5.1 | 5.5 | 5.6 | 0.3 | 0.3 | 646 |
| RB-IC-Over | 173 | 104 | 2.2 | 15.5 | 11.7 | 11.3 | 2.7 | 113 | 5.8 | 69 | 7.9 | 11.3 | 2.7 | 5.3 | 4.1 | 0.0 | 0.1 | 539 |
| RB-IC-Under | 151 | 116 | 5.6 | 19.0 | 11.5 | 20.7 | 6.7 | 104 | 6.6 | 61 | 7.8 | 20.2 | 5.6 | 5.6 | 7.0 | 0.7 | 0.4 | 549 |
| RB-OC-Over | 226 | 100 | 0.0 | 13.4 | 9.6 | 9.9 | 2.6 | 113 | 5.5 | 60 | 7.4 | 8.9 | 7.4 | 4.2 | 3.8 | 0.0 | 0.1 | 571 |
| RB-OC-Under | 384 | 144 | 4.7 | 30.5 | 15.5 | 18.6 | 3.5 | 184 | 7.3 | 81 | 11.6 | 19.9 | 4.7 | 7.1 | 7.3 | 0.4 | 0.5 | 925 |
| SB | 120 | 141 | 4.8 | 20.2 | 11.2 | 11.8 | 1.6 | 137 | 6.2 | 73 | 9.0 | 11.6 | 0.7 | 6.4 | 4.8 | 0.2 | 0.1 | 558 |
| SB-IC-Over | 113 | 166 | 7.9 | 22.6 | 12.6 | 14.2 | 1.7 | 130 | 5.3 | 76 | 8.4 | 11.3 | 0.0 | 7.5 | 6.7 | 0.2 | 0.1 | 583 |
| SB-IC-Under | 124 | 139 | 0.0 | 16.9 | 7.7 | 9.6 | 1.5 | 136 | 5.6 | 74 | 7.7 | 7.7 | 2.9 | 7.1 | 3.4 | 0.0 | 0.1 | 543 |
| SB-OC-Over | 117 | 138 | 6.9 | 22.7 | 11.7 | 14.6 | 1.8 | 141 | 7.0 | 63 | 9.6 | 14.1 | 0.0 | 4.6 | 5.1 | 0.3 | 0.1 | 558 |
| SB-OC-Under | 124 | 120 | 4.5 | 18.5 | 13.0 | 9.0 | 1.3 | 140 | 6.9 | 78 | 10.5 | 13.5 | 0.0 | 6.6 | 3.9 | 0.2 | 0.1 | 550 |



Figure 51. Average tissue PCDD/F TEQ in six sediment groups.


Figure 52. Tissue BMC PCDD/F TEQ relative abundances.


Figure 53. Tissue EC PCDD/F TEQ relative abundances.


Figure 54. Tissue PJ PCDD/F TEQ relative abundances.


Figure 55. Tissue PN PCDD/F TEQ relative abundances.


Figure 56. Tissue RB PCDD/F TEQ relative abundances.


Figure 57. Tissue SB PCDD/F TEQ relative abundances.

### 4.4. Comparison of PCDD/Fs in Tissue and Sediment.

There is a fairly strong relationship between sediment and tissue concentrations of PCDD/F TEQ for samples from Elizabeth Channel and Port Newark but the relationship is much weaker for samples from the Buttermilk Channel, Port Jersey, Raritan River, and South Brother samples where PCDD/F TEQ concentrations were lower.


Figure 58. Elizabeth Channel and Port Newark PCDD/F TEQ in sediment


Figure 59. BMC, PN, RR, and SB PCDD/F TEQ in sediment and tissue.

### 4.5 PCBs in Tissue

Table 17. Averages of the triplicate replicates from each site, $n g / g$ wet wt.

| Sed. Group | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | sum |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BMC | 0.006 | 0.250 | 2.042 | 5.392 | 5.408 | 6.396 | 3.191 | 1.102 | 0.413 | 0.211 | 24.41 |
| BMC-IC-Over | 0.005 | 0.256 | 1.066 | 3.389 | 4.145 | 5.249 | 2.331 | 0.726 | 0.352 | 0.191 | 17.71 |
| BMC-IC-Under | 0.004 | 0.197 | 1.055 | 3.244 | 4.086 | 5.041 | 2.313 | 0.713 | 0.357 | 0.195 | 17.21 |
| BMC-OC-Over | 0.007 | 0.240 | 1.719 | 4.558 | 4.895 | 5.927 | 2.808 | 0.971 | 0.375 | 0.203 | 21.70 |
| BMC-OC-Under | 0.009 | 0.307 | 4.330 | 10.376 | 8.505 | 9.368 | 5.312 | 1.998 | 0.567 | 0.256 | 41.03 |
| EC | 0.009 | 0.360 | 2.000 | 6.836 | 7.316 | 7.725 | 3.522 | 1.100 | 0.416 | 0.198 | 29.48 |
| EC-IC-Over | 0.009 | 0.486 | 2.268 | 7.281 | 8.159 | 8.430 | 3.790 | 1.122 | 0.422 | 0.199 | 32.17 |
| EC-IC-Under | 0.007 | 0.372 | 1.928 | 6.233 | 6.859 | 7.401 | 3.378 | 1.074 | 0.424 | 0.200 | 27.88 |
| EC-OC-Over | 0.005 | 0.304 | 1.655 | 5.923 | 6.250 | 7.097 | 3.173 | 0.990 | 0.404 | 0.202 | 26.00 |
| EC-OC-Under | 0.015 | 0.298 | 2.263 | 8.211 | 8.352 | 8.180 | 3.863 | 1.249 | 0.420 | 0.192 | 33.04 |
| PJ | 0.012 | 0.533 | 2.781 | 7.798 | 8.047 | 8.817 | 3.971 | 1.171 | 0.432 | 0.201 | 33.76 |
| PJ-IC-Over | 0.016 | 0.549 | 1.815 | 5.682 | 6.415 | 7.253 | 3.442 | 1.078 | 0.418 | 0.187 | 26.85 |
| PJ-IC-Under | 0.011 | 0.502 | 1.580 | 4.560 | 5.079 | 6.311 | 3.158 | 1.014 | 0.411 | 0.199 | 22.82 |
| PJ-OC-Over | 0.010 | 0.456 | 2.859 | 8.605 | 9.352 | 10.275 | 4.520 | 1.276 | 0.456 | 0.217 | 38.03 |
| PJ-OC-Under | 0.011 | 0.650 | 4.845 | 12.075 | 10.906 | 10.944 | 4.580 | 1.280 | 0.436 | 0.195 | 45.92 |
| PN | 0.007 | 0.478 | 2.790 | 9.199 | 8.736 | 8.482 | 3.771 | 1.132 | 0.411 | 0.199 | 35.21 |
| PN-IC-Over | 0.007 | 0.596 | 3.309 | 10.867 | 9.676 | 9.137 | 4.054 | 1.212 | 0.445 | 0.214 | 39.52 |
| PN-IC-Under | 0.004 | 0.371 | 2.292 | 7.791 | 7.776 | 7.664 | 3.398 | 1.040 | 0.390 | 0.189 | 30.91 |
| PN-OC-Over | 0.012 | 0.490 | 3.141 | 9.562 | 8.946 | 9.404 | 4.183 | 1.243 | 0.422 | 0.201 | 37.60 |
| PN-OC-Under | 0.008 | 0.434 | 2.377 | 8.059 | 8.584 | 8.055 | 3.632 | 1.057 | 0.363 | 0.177 | 32.75 |
| RB | 0.004 | 0.207 | 1.638 | 5.505 | 6.315 | 6.581 | 2.812 | 0.896 | 0.439 | 0.276 | 24.67 |


| RB-IC-Over | 0.004 | 0.288 | 0.865 | 3.188 | 4.204 | 6.208 | 2.897 | 0.886 | 0.393 | 0.210 | 19.14 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RB-IC-Under | 0.005 | 0.149 | 0.696 | 2.829 | 3.711 | 4.788 | 2.314 | 0.699 | 0.315 | 0.169 | 15.68 |
| RB-OC-Over | 0.002 | 0.097 | 1.346 | 4.932 | 6.458 | 6.374 | 2.719 | 0.933 | 0.494 | 0.286 | 23.64 |
| RB-OC-Under | 0.004 | 0.294 | 3.644 | 11.069 | 10.887 | 8.954 | 3.318 | 1.067 | 0.557 | 0.439 | 40.23 |
| SB | 0.006 | 0.212 | 1.154 | 3.981 | 5.787 | 7.152 | 3.482 | 0.942 | 0.408 | 0.222 | 23.35 |
| SB-IC-Over | 0.010 | 0.266 | 1.083 | 3.331 | 4.538 | 6.158 | 3.111 | 0.825 | 0.333 | 0.177 | 19.83 |
| SB-IC-Under | 0.006 | 0.215 | 1.302 | 4.091 | 6.077 | 7.054 | 3.682 | 0.922 | 0.416 | 0.217 | 23.98 |
| SB-OC-Over | 0.003 | 0.161 | 0.907 | 3.918 | 6.446 | 8.114 | 3.485 | 0.980 | 0.453 | 0.257 | 24.72 |
| SB-OC-Under | 0.004 | 0.188 | 1.347 | 4.802 | 6.505 | 7.613 | 3.771 | 1.082 | 0.453 | 0.253 | 26.02 |



Figure 60. Average tissue PCB homolog relative abundances from the six sediment groups.



Figure 62. Tissue EC PCB relative homolog abundances. A lab accident lost any possibility of detecting all three monochlorobiphenyls in one of the three EC-OC-Over samples.


Figure 63. Tissue PJ PCB relative homolog abundances.


Figure 64. Tissue PN PCB relative homolog abundances. Lab accidents lost 2-and 3-MoCB from one of the three PN-IC-Under samples and 2-, 3, and 4-MoCB from another of the PN-IC-Under samples.


Figure 65. Tissue RB PCB relative homolog abundances. A lab accident lost 9 of the 12 dichlorobiphenyls from one of the three RR-OC-Upper samples.


Figure 66. Tissue SB PCB relative homolog abundances.

Four of the sediment group averages show a similar pattern of enhanced uptake of the heavier homologs by the worms, particularly hexa- and hepta chlorobiphenyls. However, the enhancement appears to be much increased in the PJ (Port Jersey) and BMC (Buttermilk Channel) samples. At PJ the excessive uptake of the hexa- and hepta-PCBs was most pronounced in the Off-Channel samples. At BMC all four kinds of samples were about equal.

### 4.6. Comparison between PCBs in Sediment and Tissue.



Figure 67. Average tissue/sediment ratios, PCB homologs.

Table 18. $P C D D / F(\mathrm{pg} / \mathrm{g})$ and $\operatorname{PCB}(\mathrm{ng} / \mathrm{g})$ concentrations in sediment and tissue.

|  |  | PCDD/F |  | PCB |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Sed | Tissue | Sed | Tissue |
| IC | avg | 33.0 | 0.889 | 354 | 24.4 |
|  | median | 24.7 | 0.558 | 378 | 23.4 |
| OC | avg | 37.8 | 1.413 | 312 | 32.5 |
|  | median | 25.3 | 0.878 | 288 | 32.9 |
| Over | avg | 32.1 | 0.992 | 314 | 27.2 |
|  | median | 22.2 | 0.576 | 293 | 25.4 |
| Under | avg | 38.7 | 1.310 | 352 | 29.8 |
|  | median | 29.0 | 0.796 | 372 | 29.4 |

Table 19. Values are averages of the triplicate replicates from each site.

|  | PCDD/F TEQ, pg/g <br> Sed | Tissue | Tissue/Sed | PCB, ng/g <br> Sed | Tissue | Tissue/Sed |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| BMC | 15.03 | 0.50 | $3.34 \%$ | 291 | 2.44 | $0.84 \%$ |
| BMC-IC-Over | 10.96 | 0.43 | $3.92 \%$ | 271 | 1.77 | $0.65 \%$ |
| BMC-IC-Under | 13.47 | 0.37 | $2.75 \%$ | 225 | 1.72 | $0.77 \%$ |
| BMC-OC-Over | 11.30 | 0.54 | $4.78 \%$ | 242 | 2.17 | $0.90 \%$ |
| BMC-OC-Under | 24.37 | 0.67 | $2.75 \%$ | 425 | 4.10 | $0.96 \%$ |
| EC | 54.47 | 1.57 | $2.88 \%$ | 384 | 2.98 | $0.78 \%$ |
| EC-IC-Over | 39.73 | 1.37 | $3.45 \%$ | 385 | 3.22 | $0.84 \%$ |
| EC-IC-Under | 50.67 | 1.28 | $2.53 \%$ | 372 | 2.79 | $0.75 \%$ |
| EC-OC-Over | 65.28 | 1.60 | $2.45 \%$ | 368 | 2.60 | $0.71 \%$ |
| EC-OC-Under | 62.18 | 2.02 | $3.25 \%$ | 409 | 3.30 | $0.81 \%$ |
| PJ | 17.15 | 0.85 | $4.93 \%$ | 375 | 3.34 | $0.89 \%$ |
| PJ-IC-Over | 16.05 | 0.56 | $3.49 \%$ | 406 | 2.69 | $0.66 \%$ |
| PJ-IC-Under | 17.97 | 0.57 | $3.17 \%$ | 407 | 2.28 | $0.56 \%$ |
| PJ-OC-Over | 16.10 | 0.83 | $5.16 \%$ | 316 | 3.80 | $1.20 \%$ |
| PJ-OC-Under | 18.48 | 1.42 | $7.68 \%$ | 372 | 4.59 | $1.23 \%$ |
| PN | 75.41 | 2.81 | $3.73 \%$ | 338 | 3.52 | $1.04 \%$ |
| PN-IC-Over | 70.14 | 1.98 | $2.82 \%$ | 448 | 3.95 | $0.88 \%$ |
| PN-IC-Under | 75.38 | 1.99 | $2.64 \%$ | 437 | 3.09 | $0.71 \%$ |
| PN-OC-Over | 70.61 | 2.40 | $3.40 \%$ | 219 | 3.76 | $1.72 \%$ |
| PN-OC-Under | 85.52 | 4.88 | $5.71 \%$ | 250 | 3.27 | $1.31 \%$ |
| RB | 32.56 | 0.65 | $1.99 \%$ | 301 | 2.47 | $0.82 \%$ |
| RB-IC-Over | 27.29 | 0.54 | $1.98 \%$ | 261 | 1.91 | $0.73 \%$ |
| RB-IC-Under | 33.52 | 0.55 | $1.64 \%$ | 269 | 1.57 | $0.58 \%$ |
| RB-OC-Over | 26.23 | 0.57 | $2.17 \%$ | 229 | 2.36 | $1.03 \%$ |


| RB-OC-Under | 43.21 | 0.93 | $2.15 \%$ | 445 | 4.02 | $0.90 \%$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| SB | 17.64 | 0.56 | $3.16 \%$ | 314 | 2.36 | $0.75 \%$ |
| SB-IC-Over | 18.16 | 0.58 | $3.19 \%$ | 412 | 1.98 | $0.48 \%$ |
| SB-IC-Under | 22.03 | 0.54 | $2.45 \%$ | 362 | 2.40 | $0.66 \%$ |
| SB-OC-Over | 13.11 | 0.56 | $4.27 \%$ | 221 | 2.47 | $1.12 \%$ |
| SB-OC-Under | 17.27 | 0.55 | $3.18 \%$ | 260 | 2.60 | $1.00 \%$ |

The relationship between PCB concentrations in sediments and tissue is shown below.


Figure 68. Plot of total PCBs in tissue and sediment.

### 4.7. Ratios IC/OC and Over/Under for Sediment and Tissue

CARP II was designed to look for differences in PCB and PCDD/F concentrations between surface and deeper layers in sediment cores and between in-channel and off-channel sites. By looking at ratios of 2378-TCDD, total TEQ, and total PCBs we can say that generally, in-channel 2378-TCDD concentrations were lower than off-channel levels and that surface concentrations were lower than those from deeper layers.

Table 20. PCDD/F TEQ and 2378-TCDD concentration ratios in sediment and tissue.

| IC/OC | Sed. |  | Tissue |  | Over/Under | Sed. |  | Tissue |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 2378- |  | $2378-$ |  |  | $2378-$ |  | 2378- |
|  | TEQ | TCDD | TEQ | TCDD |  | TEQ | TCDD | TEQ | TCDD |
| BMC | $68 \%$ | $45 \%$ | $66 \%$ | $49 \%$ | BMC | $59 \%$ | $67 \%$ | $93 \%$ | $94 \%$ |
| BMC-Over | $97 \%$ | $90 \%$ | $80 \%$ | $75 \%$ | BMC-IC | $75 \%$ | $76 \%$ | $124 \%$ | $117 \%$ |
| BMC-Under | $55 \%$ | $32 \%$ | $55 \%$ | $37 \%$ | BMC-OC | $61 \%$ | $58 \%$ | $95 \%$ | $72 \%$ |


| EC | $73 \%$ | $65 \%$ | $74 \%$ | $71 \%$ | EC | $95 \%$ | $89 \%$ | $91 \%$ | $93 \%$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| EC-Over | $61 \%$ | $50 \%$ | $85 \%$ | $83 \%$ | EC-IC | $81 \%$ | $74 \%$ | $106 \%$ | $103 \%$ |
| EC-Under | $81 \%$ | $76 \%$ | $63 \%$ | $59 \%$ | EC-OC | $103 \%$ | $105 \%$ | $85 \%$ | $83 \%$ |
| PJ | $97 \%$ | $66 \%$ | $52 \%$ | $48 \%$ | PJ | $88 \%$ | $83 \%$ | $72 \%$ | $84 \%$ |
| PJ-Over | $100 \%$ | $63 \%$ | $68 \%$ | $56 \%$ | PJ-IC | $89 \%$ | $79 \%$ | $98 \%$ | $89 \%$ |
| PJ-Under | $97 \%$ | $70 \%$ | $40 \%$ | $40 \%$ | PJ-OC | $104 \%$ | $87 \%$ | $78 \%$ | $79 \%$ |
| PN | $90 \%$ | $75 \%$ | $54 \%$ | $50 \%$ | PN | $88 \%$ | $90 \%$ | $77 \%$ | $84 \%$ |
| PN-Over | $99 \%$ | $84 \%$ | $82 \%$ | $81 \%$ | PN-IC | $93 \%$ | $93 \%$ | $99 \%$ | $98 \%$ |
| PN-Under | $88 \%$ | $73 \%$ | $41 \%$ | $36 \%$ | PN-OC | $83 \%$ | $80 \%$ | $49 \%$ | $43 \%$ |
| RB | $89 \%$ | $94 \%$ | $69 \%$ | $50 \%$ | RB | $71 \%$ | $83 \%$ | $75 \%$ | $92 \%$ |
| RB-Over | $104 \%$ | $103 \%$ | $94 \%$ | $77 \%$ | RB-IC | $84 \%$ | $87 \%$ | $99 \%$ | $113 \%$ |
| RB-Under | $78 \%$ | $86 \%$ | $59 \%$ | $39 \%$ | RB-OC | $72 \%$ | $79 \%$ | $87 \%$ | $105 \%$ |
| SB | $130 \%$ | $113 \%$ | $102 \%$ | $97 \%$ | SB | $81 \%$ | $79 \%$ | $105 \%$ | $101 \%$ |
| SB-Over | $138 \%$ | $121 \%$ | $104 \%$ | $96 \%$ | SB-IC | $83 \%$ | $75 \%$ | $103 \%$ | $93 \%$ |
| SB-Under | $128 \%$ | $113 \%$ | $99 \%$ | $99 \%$ | SB-OC | $79 \%$ | $82 \%$ | $109 \%$ | $109 \%$ |

Figure 69 illustrates these data with the example of IC/OC ratios for 2378 -TCDD in sediment (the second data column in the above table). The second data row ( $90 \%$ ) is the ratio from Sediment Group BMC of the mean 2378-TCDD concentrations in the upper layer of the IC cores divided by the mean concentration of 2378-TCDD in the upper layer of the OC cores.


Figure 69. 2378-TCDD concentrations from sediment samples in navigational channels divided by concentrations in matched off channel areas. Bars less than $100 \%$ occur when concentrations are lower in IC samples.

Table 21. PCB concentration ratios, In-Channel/Off-Channel.

| IC/OC | Sed | Tissue | Over/Under | Sed | Tissue |
| :---: | :---: | :---: | :---: | :---: | :---: |
| BMC | 68\% | 66\% | BMC | 80\% | 68\% |
| BMC-Over | 97\% | 80\% | BMC-IC | 169\% | 103\% |
| BMC-Under | 55\% | 55\% | BMC-OC | 76\% | 73\% |
| EC | 73\% | 74\% | EC | 97\% | 94\% |
| EC-Over | 61\% | 85\% | EC-IC | 102\% | 116\% |
| EC-Under | 81\% | 63\% | EC-OC | 93\% | 80\% |
| PJ | 97\% | 52\% | PJ | 94\% | 97\% |
| PJ-Over | 100\% | 68\% | PJ-IC | 100\% | 118\% |
| PJ-Under | 97\% | 40\% | PJ-OC | 114\% | 104\% |
| PN | 90\% | 54\% | PN | 104\% | 124\% |
| PN-Over | 99\% | 82\% | PN-IC | 103\% | 130\% |
| PN-Under | 88\% | 41\% | PN-OC | 87\% | 115\% |
| RB | 89\% | 69\% | RB | 71\% | 77\% |
| RB-Over | 104\% | 94\% | RB-IC | 97\% | 130\% |
| RB-Under | 78\% | 59\% | RB-OC | 69\% | 68\% |
| SB | 130\% | 102\% | SB | 104\% | 88\% |
| SB-Over | 138\% | 104\% | SB-IC | 114\% | 90\% |
| SB-Under | 128\% | 99\% | SB-OC | 85\% | 100\% |

Generally, 2378-TCDD and to a lesser extent, total TEQ, is lower in surficial sediment and in navigational channel sediments than in deeper sediments and sediments from the off-channel cores. If we assume that surficial sediments and navigational channel sediments are younger, we may conclude that the rate of PCDD/F deposition is decreasing in some regions. The effect is stronger for IC/OC comparisons than for Over/Under comparisons. This effect is also seen in IC/OC comparisons for PCBs but less for Over/Under PCB comparisons.

It is reasonable to expect environmental 2378-TCDD concentrations to fall off faster than PCBs because there is less of it in the urban fabric.

## Chapter 5. Contribution of PCBs to total TEQ in Sediment and Tissue

The 2005 WHO list of dioxin-like toxic equivalency factors (TEFs) includes 12 PCB congeners. PCB TEFs are usually lower than those for the PCDD/Fs but PCBs are more abundant. While they are not currently considered as part of the regulatory definition of dioxin TEQ, they can contribute a significant proportion of the total TEQ in some harbor environments.

Table 22. WHO TEFs for the Co-Planar "Toxic" PCBs.

| PCB congener | TEF WHO 2005 |
| :---: | :---: |
| 3,3',4,4'-TeCB | 0.0001 |
| 3,4,4',5-TeCB | 0.0003 |
| 2,3,3',4,4'-PeCB | 0.00003 |
| 2,3,4,4',5-РеCB | 0.00003 |
| 2,3',4,4',5-РеСВ | 0.00003 |
| 2',3,4,4',5-РеСВ | 0.00003 |
| 3,3',4,4',5-РеСВ | 0.1 |
| 2,3,3',4,4',5-HxCB | 0.00003 |
| 2,3,3',4,4',5'-HxCB | 0.00003 |
| 2,3',4, 4',5,5'-HxCB | 0.00003 |
| 3,3',4, 4',5,5'-HxCB | 0.03 |
| 2,3,3',4,4',5,5'-НрСВ | 0.00003 |

Table 23. $P C D D / F$ and $P C B$ TEQs in sediment and tissue, $p g / g$.

|  | Sediment |  | PCB contribution | TISSUE |  | PCB contribution |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Row Labels | PCDD/F | PCB | to total TEQ | OCDD/F | PCB | to total TEQ |
| BMC | 15.0 | 3.9 | $21 \%$ | 0.5 | 0.3 | $35 \%$ |
| BMC-IC-Over | 11.0 | 4.0 | $26 \%$ | 0.4 | 0.3 | $37 \%$ |
| BMC-IC-Under | 13.5 | 3.3 | $20 \%$ | 0.4 | 0.2 | $39 \%$ |
| BMC-OC-Over | 11.3 | 3.4 | $23 \%$ | 0.5 | 0.3 | $34 \%$ |
| BMC-OC-Under | 24.4 | 5.1 | $17 \%$ | 0.7 | 0.3 | $33 \%$ |
| EC | 53.9 | 5.8 | $10 \%$ | 1.6 | 0.3 | $17 \%$ |
| EC-IC-Over | 39.7 | 5.5 | $12 \%$ | 1.4 | 0.4 | $23 \%$ |
| EC-IC-Under | 50.7 | 5.6 | $10 \%$ | 1.3 | 0.4 | $24 \%$ |
| EC-OC-Over | 65.3 | 6.1 | $9 \%$ | 1.6 | 0.3 | $14 \%$ |
| EC-OC-Under | 62.2 | 6.0 | $9 \%$ | 2.0 | 0.3 | $12 \%$ |
| PJ | 17.3 | 4.7 | $21 \%$ | 0.8 | 0.5 | $37 \%$ |
| PJ-IC-Over | 16.1 | 4.3 | $21 \%$ | 0.6 | 0.6 | $50 \%$ |
| PJ-IC-Under | 18.0 | 4.5 | $20 \%$ | 0.6 | 0.5 | $46 \%$ |
| PJ-OC-Over | 16.1 | 3.8 | $19 \%$ | 0.8 | 0.5 | $36 \%$ |
| PJ-OC-Under | 18.5 | 5.4 | $23 \%$ | 1.4 | 0.5 | $25 \%$ |
| PN | 75.4 | 4.9 | $6 \%$ | 2.4 | 0.3 | $12 \%$ |
| PN-IC-Over | 70.1 | 5.9 | $8 \%$ | 2.0 | 0.4 | $17 \%$ |
| PN-IC-Under | 75.4 | 5.6 | $7 \%$ | 2.0 | 0.3 | $12 \%$ |
| PN-OC-Over | 70.6 | 2.5 | $3 \%$ | 2.4 | 0.3 | $11 \%$ |
| PN-OC-Under | 85.5 | 3.2 | $4 \%$ | 4.9 | 0.3 | $6 \%$ |
| RB | 32.6 | 4.9 | $13 \%$ | 0.7 | 0.4 | $37 \%$ |


| RB-IC-Over | 27.3 | 4.7 | $15 \%$ | 0.5 | 0.4 | $45 \%$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| RB-IC-Under | 33.5 | 4.8 | $12 \%$ | 0.6 | 0.3 | $38 \%$ |
| RB-OC-Over | 26.2 | 3.9 | $13 \%$ | 0.6 | 0.3 | $36 \%$ |
| RB-OC-Under | 43.2 | 6.3 | $13 \%$ | 0.9 | 0.4 | $32 \%$ |
| SB | 17.6 | 5.7 | $25 \%$ | 0.6 | 0.4 | $39 \%$ |
| SB-IC-Over | 18.2 | 5.1 | $22 \%$ | 0.6 | 0.4 | $38 \%$ |
| SB-IC-Under | 22.0 | 6.4 | $22 \%$ | 0.5 | 0.3 | $35 \%$ |
| SB-OC-Over | 13.1 | 4.1 | $24 \%$ | 0.6 | 0.4 | $42 \%$ |
| SB-OC-Under | 17.3 | 7.0 | $29 \%$ | 0.6 | 0.4 | $42 \%$ |

## Chapter 6. PCDD/F and PCB Data Quality Review

### 6.1. Completeness

The completeness of the principal study components is shown below.

Table 24. Sampling completeness for water and sediment cores for PCDD/Fs and PCBs.

| study | analyte | expected | achieved | $\%$ completeness |
| :--- | :--- | :---: | :---: | :---: |
| water | PCDD/Fs | 34 | 32 | $94 \%$ |
| water | PCBs | 68 | 66 | $97 \%$ |
| sediment core | PCDD/Fs | 72 | 68 | $94 \%$ |
| Sediment core | PCBs | 72 | 68 | $94 \%$ |

In five samples some of the mono- or dichlorobiphenyls were lost due to lab accidents, flagged " NQ ". These are noted in the discussion below. They do not materially impact total PCB concentrations from the samples.

### 6.2. Non-detections

The CARP Management team developed a procedure for handling non-detections:
[conc_found] = concentration found
[DL] = sample specific detection limit
[ratio_conc] = the frequency of detection of an analyte divided by the number of analyses for that analyte by medium (sediment, TOPS cartridge, tissue, water) and by site type (sediment, ambient, HOT, Poughkeepsie, storm water)

If [conc_found]>[DL], [conc_found], then use [conc_found] otherwise,
If [ratio_conc] $>0.7$, then use [DL]/2 other wise, 0 .

### 6.3. Flagged data

Contaminant analysis for PCDD/F and PCBs was performed by the contract lab AXYS. Axys reported seven distinct data quality flags, other than blanks (null) indicating no quality issues. The database contains the following data quality flags:
$B$ - analyte found in lab blank.
C - coelution, applies only to PCB samples.
D - sample required dilution.
$J$ - reported value is less than the method detection level but above the sample detection level.
U - analyte less than sample detection level and reported as null.
K - all method criteria were not satisfied, reported value is an estimated maximum.
G - lock mass disturbances ${ }^{2}$

The percent frequency of data quality flags by medium and sample type are shown in Table 24. Samples may have multiple flags so these percents add up to more than $100 \%$.

Table 24. Frequency of flagged analyses.

|  |  | null, C or D | ${ }^{*} \mathbf{U}^{*}$ | ${ }^{*} \mathbf{B}^{*}$ | ${ }^{*} \mathbf{J}^{*}$ | ${ }^{*} \mathbf{K}^{*}$ or ${ }^{*} \mathbf{G}^{*}$ |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| Cartridge Water | DIOX/F | $0 \%$ | $43 \%$ | $27 \%$ | $49 \%$ | $16 \%$ |
| Sediment | DIOX/F | $41 \%$ | $1 \%$ | $45 \%$ | $18 \%$ | $2 \%$ |
| Susp_Sed | DIOX/F | $25 \%$ | $10 \%$ | $39 \%$ | $39 \%$ | $14 \%$ |
| Tissue | DIOX/F | $12 \%$ | $12 \%$ | $18 \%$ | $68 \%$ | $20 \%$ |
| Sediment | PCB | $45 \%$ | $9 \%$ | $39 \%$ | $5 \%$ | $5 \%$ |
| Susp_Sed | PCB | $22 \%$ | $17 \%$ | $43 \%$ | $26 \%$ | $13 \%$ |
| Tissue | PCB | $48 \%$ | $10 \%$ | $37 \%$ | $2 \%$ | $6 \%$ |
| Water | PCB | $10 \%$ | $25 \%$ | $37 \%$ | $40 \%$ | $17 \%$ |

The significance of the lab blanks (B) can be assessed by looking at how much the reported concentration exceeds the blank. In the following table we see that $94 \%$ of the tissue samples for PCDD/Fs exceeded their associated blanks by 10 times and $98 \%$ were twice the blank. Cartridge water samples were more problematic.

[^1]In the subsequent data tables all PCDD/F and PCB data are blank corrected where the lab blank for each work group is subtracted from the reported value. Where the blank was reported as non-detect, the subtracted value was 0 .

Table 25. Frequency of data values exceeding the associated blanks by 10 and 2 times.

|  |  | \%>10X | \%>2X |
| :--- | :--- | :---: | :---: |
| Cartridge Water | DIOX/F | $22 \%$ | $59 \%$ |
| Sediment | DIOX/F | $94 \%$ | $98 \%$ |
| Susp_Sed | DIOX/F | $67 \%$ | $86 \%$ |
| Tissue | DIOX/F | $50 \%$ | $74 \%$ |
| Sediment | PCB | $100 \%$ | $100 \%$ |
| Susp_Sed | PCB | $77 \%$ | $89 \%$ |
| Tissue | PCB | $93 \%$ | $99 \%$ |
| Water | PCB | $66 \%$ | $90 \%$ |

Four PCDD/F congeners accounted for $66 \%$ of total TEQ across all media. The data quality flags for these compounds are shown below by medium.

Table 26. Frequency of data flags on the PCB and PCDD/F congeners most critical to total TEQ.

| medium | Order | analyte | null or D | *U* | *B* | *J* | *K* or *G* | \# samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cart. Water | 1 | 2,3,7,8-TCDD |  | 4 |  | 4 | 4 | 8 |
| Cart. Water | 2 | 1,2,3,7,8-PECDD |  | 5 |  | 3 |  | 8 |
| Cart. Water | 8 | 2,3,7,8-TCDF |  | 4 |  | 4 | 3 | 8 |
| Cart Water | 10 | 2,3,4,7,8-PECDF |  | 5 |  | 3 | 2 | 8 |
| Sediment | 1 | 2,3,7,8-TCDD | 39 |  | 35 |  | 1 | 74 |
| Sediment | 2 | 1,2,3,7,8-PECDD | 29 | 1 |  | 43 | 3 | 74 |
| Sediment | 8 | 2,3,7,8-TCDF | 73 |  | 1 |  |  | 74 |
| Sediment | 10 | 2,3,4,7,8-PECDF | 22 |  | 51 | 6 |  | 74 |
| Susp_Sed | 1 | 2,3,7,8-TCDD | 9 | 6 | 20 | 10 | 16 | 42 |
| Susp_Sed | 2 | 1,2,3,7,8-PECDD | 9 | 9 | 2 | 23 | 14 | 42 |
| Susp_Sed | 8 | 2,3,7,8-TCDF | 36 | 5 |  |  | 2 | 42 |
| Susp_Sed | 10 | 2,3,4,7,8-PECDF | 9 | 4 | 2 | 27 | 6 | 42 |
| Tissue | 1 | 2,3,7,8-TCDD | 32 | 1 |  | 31 | 14 | 70 |
| Tissue | 2 | 1,2,3,7,8-PECDD |  | 3 | 16 | 66 | 30 | 70 |
| Tissue | 8 | 2,3,7,8-TCDF | 61 |  |  |  | 9 | 70 |
| Tissue | 10 | 2,3,4,7,8-PECDF | 1 |  |  | 69 | 8 | 70 |

The impact of K-flagging on total TEQ was evaluated by calculating the ratio of samples with K and G flagged instances over uncensored samples.

Table 27. Impact of $k$-flagging on total TEQ, by medium.

|  | Susp_Sed | Sediment | Tissue |
| :---: | :---: | :---: | :---: |
| TOPS Samples |  |  |  |
| Ambient | $92 \%$ |  |  |
| HOT | $83 \%$ |  |  |
| Poughkeepsie | $93 \%$ |  |  |
| Storm Water | $91 \%$ |  |  |
| Sediment |  |  |  |
| BMC |  | $95 \%$ | $82 \%$ |
| EC |  | $100 \%$ | $86 \%$ |
| PJ |  | $99 \%$ | $63 \%$ |
| PN |  | $100 \%$ | $97 \%$ |
| RR |  | $100 \%$ | $72 \%$ |
| SB |  | $98 \%$ | $80 \%$ |

Generally, the significance of K-flagged instances in calculating TEQ was greatest in tissue samples. Some areas were more problematic than others. The magnitude of error from $K$ and $G$ flagging was assessed for the PCBs in a similar manner. For example, the average PCB concentration from ambient water samples would be reduced by $15 \%$ if all the observations with $K$ or $G$ flags were set to zero due to their uncertainty.

Table 28. Impact of $k$-flagging on PCB concentrations by medium. 6

|  | Susp_Sed | Water | Sediment | Tissue |
| :---: | :---: | :---: | :---: | :---: |
| TOPS Samples |  |  |  |  |
| Ambient | $91 \%$ | $85 \%$ |  |  |
| HOT | $97 \%$ | $98 \%$ |  |  |
| Poughkeepsie | $95 \%$ | $91 \%$ |  |  |
| Storm Water | $96 \%$ | $92 \%$ |  |  |
| Sediment |  |  |  |  |
| BMC |  |  | $94 \%$ | $88 \%$ |
| EC |  |  | $95 \%$ | $92 \%$ |
| PJ |  |  | $91 \%$ | $87 \%$ |
| PN |  |  | $87 \%$ | $91 \%$ |
| RB |  |  | $92 \%$ | $87 \%$ |
| SB |  |  | $91 \%$ | $87 \%$ |

### 6.4. A measure of quantitative reliability

Quantitation of chemical targets are more certain when they are well above the detection limits. As a rule of thumb, good quantitations are ten or more times greater than the detection level. To simplify an examination of the adequacy of the project's sensitivity, we look at the average ratios of observed/DL for the four homologs. Detection of these key congeners was uniformly strong in sediment samples and of variable quality among the TOPS and Tissue samples.

Table 29. Average ratios, reported analyte concentrations/sample specific detection limits of four PCB homologs.

|  | Susp_Sed |  |  |  | Sediment |  |  |  | TISSUE |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Homologs | 1 | 2 | 8 | 10 | 1 | 2 | 8 | 10 | 1 | 2 | 8 | 10 |
| TOPS |  |  |  |  |  |  |  |  |  |  |  |  |
| Samples |  |  |  |  |  |  |  |  |  |  |  |  |
| Ambient | 18.5 | 2.1 | 35.4 | 6.3 |  |  |  |  |  |  |  |  |
| Head_of_Tide | 3.4 | 7.9 | 17.7 | 5.4 |  |  |  |  |  |  |  |  |
| Poughkeepsie | 4.1 | 6.4 | 53.0 | 17.9 |  |  |  |  |  |  |  |  |
| Storm_Water | 8.7 | 26.3 | 32.0 | 6.6 |  |  |  |  |  |  |  |  |
| Sediment |  |  |  |  |  |  |  |  |  |  |  |  |
| BMC |  |  |  |  | 28.4 | 11.5 | 147.7 | 25.9 | 2.6 | 2.3 | 16.0 | 3.0 |
| EC |  |  |  |  | 175.2 | 14.3 | 255.8 | 37.6 | 16.3 | 2.3 | 19.5 | 4.9 |
| PJ |  |  |  |  | 26.7 | 10.2 | 125.6 | 26.5 | 4.9 | 3.1 | 22.0 | 5.6 |
| PNC |  |  |  |  | 222.9 | 16.4 | 257.7 | 30.9 | 23.5 | 2.2 | 28.6 | 4.1 |
| RR |  |  |  |  | 55.0 | 22.6 | 251.3 | 51.5 | 4.6 | 2.0 | 21.5 | 3.6 |
| SB |  |  |  |  | 22.1 | 18.9 | 194.1 | 47.3 | 1.9 | 2.1 | 20.8 | 3.5 |

For the PCBs the average ratios of observed/detection limits less than 10X show relatively poor quantitations for mono, nona- and octa-CBs in $15 \%$ of water samples. Six percent of suspended sediment samples, mostly from HOT sites, had ratios less than 10X. One percent of tissue sample mono and diCBs had conc/DL ratios less than 10X. Detection limits were never a problem with sediments.

No field blank samples were taken for sediments or tissues. Four were made for the TOPS samples. Field blanks are compared against samples by pg/recovered, not calculated concentrations. The highest field blank (total TEQ of 0.8116 ) was 7 times lower than the lowest sample ( $5.466 \mathrm{pg} / \mathrm{sample}$, from the Lower Bay). Contamination of the TOPS or the cartridge filters do not appear to have been significant. The field blanks were not significantly greater than the lab blanks.

## Chapter 7. Data Management and Data Sharing

The data collected under CARP II are freely and publicly available through the NOAA Data portal. The CARP II researcher teams strove to ensure that the data developed under the project was
complete, accurate and useful. However, given the complexity of the data collection and analysis processes, users should make the own determination of the quality, accuracy or completeness of the data and the fitness of use for a particular purpose.

## NOAA DIVER Data Portal

The Data collected under CARP II are available through the NOAA data portal. https://www.diver.orr.noaa.gov/

1. Choose Hudson River, Keyword Search: CARP II (Figure 1)

f DIVER EXPLORER HUDSON RIVER WATERSHED


[^0]:    ${ }^{1}$ Dredged material testing methods for were modified to use less sediment, keeping the sediment to water ratio consistent with the standard protocols.

[^1]:    ${ }^{2}$ The "lock mass" is a compound that is monitored to demonstrate that the instrument sees the ions that are being monitored accurately. Things that interfere with the instrument's ability to see this ion show a deflection in the lock mass channel. When the instrument can see the lock mass target it shows a straight-ish line in the chromatogram of the lockmass channel. But when the instrument's ability to see the compound is interfered with (usually caused by matrix interferences) you see a defection in that straight-ish line. If the deflection is at the same retention time as a target but <20\% it's considered tolerable but when it is deflected greater than $20 \%$, but not horrible we report the result and flag it with a " G ". Lockmass interference can be so severe that the result is not reportable- then we have to flag as NQ (not quantifiable).

